# Case Studies of Advances in Bioremediation of Organics: Part 2

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# 4 Case Studies

- Combined anaerobic and aerobic bioaugmentation
	- groundwater-surface-water interface
	- bioaugmentation on activated carbon
- TCE in fractured rock— permeable units
	- microbial community analyses to track bioaugmentation effects, DCE/VC stall
	- TCE in fractured rock— rock matrix
		- borehole test to quantify diffusion and degradation
	- Anaerobic PFAS biodegradation with chlorinated solvent co-contaminants









1. Coupled Anaerobic – Aerobic Biodegradation Dual-Biofilm Reactive Barrier for Treatment of Chlorinated Benzenes at Anaerobic-Aerobic Interfaces (NIEHS R01ES024279 & EPA Region III)

- Anaerobic reductive dechlorination
	- Highly chlorinated CBs thermodynamically favorable
	- Toxic daughter products remain
	- Mineralization possible, but MCB stall common
- Aerobic oxidation– oxygen-mediated process
	- Less chlorinated CBs favorable thermodynamically (<= TeCB)
	- Complete mineralization

#### **Anaerobic reductive dechlorination**





Fields and Sierra-Alvarez. 2008. Biodegradation



- Chemical plant 1966-2002
- 1986 tank failure released >500,000 gal 14DCB and TCBs that flowed over wetland
- Superfund site 1987; EPA 2002
- Underlying DNAPL-contaminated aquifer
- Abuts Red Lion Creek, part of Delaware River watershed



# Permeable Reactive Bio-barrier



• Intercept discharge and treat

### Bench-scale column studies: Simulate anoxic-oxic interface



*Conceptual model*



- Isolate biodegradation activity – no GAC amendment
- Defined, constant media composition and flux
- Inoculate with microbial enrichments:
	- Anaerobic WBC-2
	- Aerobic enrichment from site water
- Model contaminant:
	- 1,2,4-trichlorobenzene



*(Chow et al., 2020, J Contaminant Hydrology, https://doi.org/10.1016/j.jconhyd.2020.103639)*

### Stratification of Putative CB Degraders along oxic-anoxic interface Next generation 16s rRNA sequencing [Relative abundance × total 16S gene copy #]



Anaerobic Dechlorinator Dehalobacter (Clostridia) **Aerobic Oxidizers** Pandoraea (Betaproteobacteria) Unspecified Sphingomonadaceae (Alphaproteobacter Low enrichment (<1%) in Xanthobacter (Alphaproteobacteria) **MCB**  $1.2$ -DCB  $1,3-DCB$  $1,4-DCB$ 

1,2,4-TCB

Chloride

Sulfate

- 
- *Dehalobacter* enriched in biofilm as anaerobic dechlorinator
	- High enrichment (up to 50%) in sediment column
	- sand column
- Shift in degrading populations across interface
- Abundances corresponded to more favorable degradation outcomes
	- Sediment reductive dechlorination
	- Sand aerobic degradation

#### *Sulfate-reducing conditions*



 $\Omega$ 





#### Alternative Redox Conditionsyear-long experiment



- $SO_4^2$  negatively impacted reductive dechlorination; reduced S<sup>-</sup> downgradient negatively impacts aerobic degradation
- $NO<sub>3</sub>$  negatively impacted reductive dechlorination; enhanced aerobic degradation, serving as sink for competing edonors

#### *Nitrate-reducing conditions*





#### Pilot Test Preparation: GAC Effect on Culture Growth



Significant increase in Dehalococcoidales on GAC.

Dehalococcoidaceae;g Dehalococcoides Dehalococcoidaceae;g Dehalogenimonas



Anaerobic Culture WBC-2 Aerobic Culture (Native, 15B)

#### *Field Pilot Test*







# Pilot Test: Chloride

Chloride increased by **factor of 2 to 5**  in reactive zone, showing CB degradation was a major removal mechanism.



### Field-Incubated and Lab-Incubated GAC: Microbial Community



- Burkholderiales- aerobic degraders in 15B culture and in GAC in lab and field
- Desulfuromonadales- anaerobes in WBC-2 culture that increased in abundance on field incubated GAC
- Dehalococcoidia- consistently present in low percent abundance in field incubated GAC

*All data are provisional.*

Sediment Samples Taxonomy – Percent abundance by depth for select orders (control vs. reactive plots)



- Aerobes remained relatively abundant at depth in the reactive barriers compared to the controls.
- Methanogens and anaerobic dechlorinators increased in abundance with depth in control and reactive barrier plots.

*Anaerobes and aerobes overlap at all depths.*

*All data are provisional.*

# In Situ Microcosms

**Bio-Traps** (Microbial Insights), with and without Biosep beads pre-loaded with <sup>13</sup>Clabeled monochlorobenzene.

- Measure incorporation of  $^{13}$ C in CO<sub>2</sub> and PLFA
- Analysis of anaerobic and aerobic functional genes by advanced qPCR to relate microbial presence to degradation ability

#### *Concurrent microbial and isotopic data to verify biodegradation activity.*



*Is biodegradation in the reactive barriers enhanced compared to the control sediment areas, and does aerobic and anaerobic biodegradation co-occur?*

### In Situ Microcosms and Advanced qPCR: Aerobes  $\begin{array}{ccc} \text{Aerobes} & \text{I m} \ & & \text{I m} \end{array}$



- Aerobic oxidation indicators higher in reactive barrier at site 135 compared to site 8 reactive barrier or controls.
- Agrees with next generation sequencing results.





Reactive



# In Situ Microcosms and Advanced qPCR: Anaerobes



- DSM (*Desulfuromonas*) was the only reductive dechlorination indicator that was consistently higher in the reactive barrier plots than controls.
- Indicators of both aerobic oxidation activity and anaerobic reduction present in all ISMs.



**DHBt** *Dehalobacter* **spp.** DCM *Dehalobacter* DCM DHG *Dehalogenimonas* spp. DSB *Desulfitobacterium* spp. **DECO** *Dehalobium chlorocoercia* DSM *Desulfuromonas* spp.



### In Situ Microcosms: 13C from Labeled MCB in Biomass



- **High 13C uptake in biomass** (PLFA) in the reactive barrier at site 135 indicates high aerobic oxidation of MCB.
- Agrees with the observed higher abundance of aerobic oxidizers and functional genes at site 135 compared to site 8.

### In Situ Microcosms:  $^{13}$ C from Labeled MCB in CO<sub>2</sub>



- Incorporation of <sup>13</sup>C in CO<sub>2</sub> was **high in both reactive barriers** and low in the controls, verifying complete enhanced biodegradation in the reactive barriers.
- Complete degradation to  $CO<sub>2</sub>$  is  $\sim$  equal in the two reactive barriers, despite the lower use of MCB as growth substrate at site 8. Indicates a combination of anaerobic  $(^{13}C$  for energy) and aerobic biodegradation processes in the reactive barrier.

2. TCE in Fractured Rock- Permeable Units Microbial community dynamics with next-generation sequencing to monitor bioremediation

- Examined archived DNA from prior bioaugmentation pilot (SERDP ER-1555) with next generation sequencing
- Metagenomics to better understand bioremediation effects and "DCE stall"





#### Former Naval Air Warfare Center (NAWC) West Trenton, NJ

- Focus site for USGS research on fracturedrock contamination
- Several SERDP and ESTCP studies
- *One-time injection of EOS and KB-1 consortium*
- *2008-2015*





Predominant taxa with time along flowpath

- Microbial shifts were observed in downgradient wells 73BR -D2 and 71BR -C, but not in the pumping well, 15BR.
- An increased predominance of *Geobacter* spp. (partial dechlorinators) marked the microbial effects of bioaugmentation.

#### Shared OTUs



*Bioaugmented populations out-competed by native bacteria*

# 3. TCE in Rock Matrix-Borehole Test (NAWC)

#### Post-drilling conditions



- Permeable fractures feed TCE into wellbore water
- TCE diffuses into low-permeability matrix post-drilling
- Concentration history of TCE, DCE,



field method for diffusion & sorption coefficients & cVOC reaction rates in low-permeability strata

VC used as boundary condition<br>SERDP ER-2533 & USGS Toxic Substances Hydrology Program



**Microbial samplers.** made with clean Ottawa sand inside stainless steel mesh. placed in test interval between packers and left for duration of tracer test.

# TCE Degradation in Microcosms-Borehole water and *in situ* incubated sand



- **TCE degradation rates same with and without WBC-2 addition**
- Addition of WBC-2 did result in 12DCE and VC degradation; indicates lack of necessary microbial species/densities for complete dechlorination.

### Borehole Test- Measured and Simulated cVOC **Concentrations**

**Borehole Matrix** packer · Diffusion • Mass Balance Br, TCFE, TCE, cDCE, VC, ADP • Biodegradation **TCE TCFE** • Sorption ÷ TCFE, TCE, cDCE, VC **cDCE DCFE VC** • Abiotic Degradation ?? **TCE → Abiotic Degradation CE Products (ADP)** packer







### Comparison of Lab and Field Test Rates



# Impact of Biodegradation in Borehole



- Biodegradation of TCE to cDCE in borehole is significant under experiment (no flow) conditions
	- maintains low TCE concentration compared to matrix
	- biodegradation occurs at the interface, not in the matrix
	- cDCE to VC degradation rate increased in later part of test
- Sorption is significant
	- cVOC mass stored mostly within a few cm's of the borehole wall.

Fracture:  $(196)$ 

**Rock Matrix**  $(6,854)$ 

• Abiotic degradation not detected

*Previous study showed 97 % of TCE mass is in rock matrix (Goode et al. 2014)*

### Family level in microcosms (sand sampler from borehole)

- *MA-28-198C in Desulfuromonadales order, close to Geobacter*
- *Increase in abundance delayed at high conc.*



- $WCHB1-32$  fa
- Cloacimonetes Incertae Sedis Unknown Order Unknown Family
- $\blacksquare$  Rhodobacteraceae
- Rhodocyclaceae
- Desulfomicrobiaceae
- Syntrophobacteraceae
- Synergistaceae
- Candidatus Berkelbacteria fa
- Peptococcaceae
- Comamonadaceae
- Desulfobacteraceae
- Geobacteraceae
- Helicobacteraceae
- Kosmotogaceae
- Anaerolineaceae
- Veillonellaceae
- Methylophilaceae
- Desulfobulbaceae
- MA-28-198C
- Spirochaetaceae

# 4. Anaerobic PFAS Biodegradation

- $\blacktriangleright$  Biodegradation studies using sediment in groundwater discharge areas- diversity of microbial species
- $\blacktriangleright$  Initial focus on co-contaminant effects with cVOCs and added organohalide-respiring bacteria
- $\blacktriangleright$  Sediment samples for initial test collected at Ft. Drum, NY (by US Army Corps of Engineers, Baltimore)



*<sup>(</sup>modified from Lorah et al., 2005)*

Overlap in species capable of reductive dechlorination and reductive defluorination?



## Microbial Reductive Defluorination

 $\blacktriangleright$  Generally believed not possible for the perfluorinated compounds, although is thermodynamically feasible.

- $\blacktriangleright$  Highly chlorinated and brominated organics that initially were believed to be recalcitrant are now known to undergo reductive dehalogenation.
- ▶ An association between Chloroflexi abundance and PFAS contamination was shown in soil (Chen et al, 2019) and river sediments (Bao et al., 2018; PFOS in particular)
- $\blacktriangleright$  Recent identification by Huang and Jaffé (2019) of defluorination by *Acidimicrobium* sp. strain A6, a natural microbe that also dechlorinates PCE, TCE.

# PFAS Microcosms— PFOS Removal

- **PFOS removal in two** treatments with added cVOCs— site sediment with and without WBC-2
- $\blacktriangleright$  PFOS removal greatest with added WBC-2 and cVOCs
- ▶ 25 and 45% removal (after account for loss in control)
- PFOA and 6:2 FtS added in the same microcosms did not show removal





### PFAS Micrososm: cVOC degradation



- Fast cVOC degradation in WBC-2 bioaugmented microcosm and low daughter product accumulation.
- Slower cVOC degradation and greater daughter product detections in the non-bioaugmeted site sediment.

*All data are provisional.*

# PFOS and cVOCs Removal Rates





*organofluorine compounds, measured by Total Oxidizable Precursor (TOP) Assay on day 1 and day 45 microcosm samples.*

- $\blacktriangleright$  Link indicated between cVOC degraders and PFOS degraders.
- $\blacktriangleright$  Microbes involved in PFOS transformation in native sediment possibly associated with alkane degradation pathway? *Removal of total*



*All data are provisional.*



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#### s

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#### 2. TCE in Fractured Rock-Permeable

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#### *in bold: contributed figures or slides*