



Although I'm sure that some of you have these rules memorized from previous CLU-IN events, let's run through them quickly for our new participants.

Please mute your phone lines during the seminar to minimize disruption and background noise. If you do not have a mute button, press *6 to mute #6 to unmute your lines at anytime. Also, please do NOT put this call on hold as this may bring delightful, but unwanted background music over the lines and interupt the seminar.

You should note that throughout the seminar, we will ask for your feedback. You do not need to wait for Q&A breaks to ask questions or provide comments. To submit comments/questions and report technical problems, please use the ? Icon at the top of your screen. You can move forward/backward in the slides by using the single arrow buttons (left moves back 1 slide, right moves advances 1 slide). The double arrowed buttons will take you to 1st and last slides respectively. You may also advance to any slide using the numbered links that appear on the left side of your screen. The button with a house icon will take you back to main seminar page which displays our agenda, speaker information, links to the slides and additional resources. Lastly, the button with a computer disc can be used to download and save today's presentation materials.

With that, please move to slide 3.











Does anyone feel that having a little fun during the day is too juvenile, like working for reward by answering questions in quizzes?

We try to liven up a dry, dry...dry, dry topic and a long course to keep you interested and awake.

Speak now or forever hold your peace.



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Speak now or forever hold your peace.







Incremental vs. Composite Sampling Basic Differences

- Goal of incremental sampling is to find an average concentration over some defined area (a DU).
 - Uses 3 or more independent replicates to obtain a UCL
- Goal of composite sampling is to gain information about the spatial distribution of contaminants.
 - <u>Accurate</u> estimate of the concentration mean within the given area not required
 - Typically not use replicates over the composite area
 - Suppresses short-scale heterogeneity for a "single" sample & improves representativeness
- Composite uses fewer increments per sample than incremental sampling

12













Emphasize that this is cartoon whose only purpose is to illustrate the sample support concept.



Emphasize that this is cartoon whose only purpose is to illustrate the sample support concept.

#1 would represent the layer of interest better if the depth of the sample was the same as the depth of the layer of interst.

<section-header><text><text><text><text><page-footer>



8/5/98 Preamble page 35409; Codified 63FR 35465 & 35466

1.5 m x 1.5 meters = 4.5 ft x 4.5 ft = 20.25 sq ft

1985 PCB guidance on EPA.gov http://www.epa.gov/epawaste/hazard/tsd/ pcbs/pubs/subpartmopr.pdf



Animated slide for presentation







8/5/98 Preamble page 35409; Codified 63FR 35465 & 35466

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1985 PCB guidance on EPA.gov http://www.epa.gov/epawaste/hazard/tsd/ pcbs/pubs/subpartmopr.pdf

Spatial Scales of Data Variability			
Data Variability: variatio magnitude of a popula	ns in the m tion charac	easurec cteristic	ł
Micro-scale, within sample jar: measured by lab duplicates (2 analytical subsamples taken from same sample jar) Pb: ICP lab duplicates			
Lab dup examples from actual site data. Population characteristic being measured = Pb conc	original	dup	
	1038	688	
	332	412	
	874	2187	
	34	49	
	248	223	
	69	45	
2/16/2012 Clu-In Incremental-Comp	oosite Webinar Module 1	.2	26

















"Given the average concentration in soil, the iron in a cubic yard of soil is capable of adsorbing from 0.5 to 5 pounds of soluble metals as cations, anionic complexes, or a similar amount of organic[s]." (Vance, 1994). [Reference = David B. Vance. National Environmental Journal. May/June. 1994 Vol.4 No. 3 page 24-25.

Mass of As in 1 ug soil with 5000 ppm conc (0.5%)= 5 ng If that 5 ng As dropped into 1 g soil = 5 ppb If that 5 ng As dropped into 1 kg soil = 5 ppt 5 ng dropped into 0.5g soil (ICP) = 10 ppb 5 ng dropped into 2 g soil (ICP) = 2.5 ppb

Clays Have a Layered Structure and a Very Small Particle Size





Cartoon of smectite plates with an embedded positively charged metallic ion (such as calcium, purple ball). The plate structure increases the surface area and attractive forces available to bind a contaminant molecule (DD = dibenzop-dioxin; red balls represent oxygen atoms) (Source: SRP Research Brief 183, 3/3/10, page 2).


Composition & Particle Size Create Within- Sample (Micro-Scale) Heterogeneity			
Adapted from ITRC (2003)	Firing Range Soil Grain Size (Std Sieve Mesh Size)	Pb Concentration in fraction by AA (mg/kg)	
	Greater than 3/8" (0.375")	10	
	Between 3/8" and 4-mesh	50	
	Between 4- and 10-mesh	108	
	Between 10- and 50-mesh	165	
	Between 50- and 200-mesh	836	
	Less than 200-mesh	1,970	
	Bulk Total	927 (wt-averaged)	
2/16/2012 Clu-In Incremental-Composite Webinar Module 1.2			3

Data adapted from the Interstate Technology and Regulatory Counsel (ITRC). 2003. Characterization and Remediation of Soils at Closed Small Arms Firing Ranges. January. Available on-line at http://www.itrcweb.org/SMART-1.pdf

Arms Firing Ranges. January. Available on-line at http://www.itrcweb.org/SMART-1.pdf The results of this study show how different particle sizes within the same jar of soil have different lead (Pb) concentrations. We call this "within-sample" or "micro-scale" soil heterogeneity because different concentrations of analyte occur on very small spatial levels within in a single jar of soil. Although the soil may look "homogenized," it really isn't as long as different particles sizes exist in the sample jar. This would not matter IF the entire volume of soil in the jar was analyzed all at once. Analyzing the whole sample gives you the true concentration of the jar contents. However, jars usually contain 100 grams or more of soil. Common analytical methods for Pb (and other metals) use between 0.5 and 2 grams of soil for the analysis, depending on the lab's SOP. So the analytical sample is much, much smaller than the mass of soil in the jar. For this study, a large soil sample was taken from a firing range with Pb contamination. The soil sample was dried and clods were broken apart, but no grinding was performed. Visible fragments of Pb bullets were removed. The soil was then sieved into different-size fractions. The 6 particle size fractions that resulted are provided above. Particle size gets smaller as the mesh size increases. Each particle-size fraction was analyzed for Pb separately by atomic absorption (AA), a routine laboratory method for analyzing metals. An obvious trend exists for this site's soil: the Pb concentration in a particle size fraction increases as the particle size each aparticle was particle with the particle size, the more surface area is available to adsorb contaminants like Pb. And the smallest fraction is more likely to have particle size, the more surface area is available to adsorb contaminants like Pb. And the smallest fraction is more likely to have particle size fraction the soil constituents that have properties that attract and holds on to positively-charged metal ions. Over time,

Particle size effects on analytical results have ramifications for the sampling & analysis of soil. When soil is shipped to a lab, motions in transit cause a segregation of particle sizes within jars. When a sample jar arrives at the lab, larger particles are typically sitting on top, and smaller particles have moved toward the bottom. If a technician were to sample a jar by unscrewing the cap and simply scooping a subsample off the top, the Pb result would likely be a lot lower than the true Pb concentration for the whole jar of soil.

As mentioned above, metals analysis for soil typically involves digesting a very small mass, around 1 gram. So another variable that can affect the concentration of the analysis for soil typically involves digesting a very small mass, around 1 gram. So another variable that can affect gram. A variety of utensils of varying sizes and configurations can be used to scoop up small amounts of soil and ferry it from the jar to the weigh boat that sits on the balance. There is no standardization of what utensil should be used. Even within the same lab, different technicians may use different scoops. A larger, spoon-shaped utensil will retain the larger particles (which provide mass, but little Pb), but those particles could easily roll of a flat spatula or a much smaller scoop. Thus a larger bowl-shaped utensil will select FOR larger particles, whereas a flat or very small scoop surface will select AGAINST larger particle sizes.

Another variable is related to the motions the technician makes while weighing out the analytical sample. Say the target mass for an analytical sample was 1 gram. Weighing out samples takes time, and technicians are always under pressure to maintain high sample throughput. So the fewer scoops into the weigh boat needed to get close to 1 gram, the more samples a technician can process. So naturally, the technician will make the 1st scoop out of the jar larger to try to get close to 1 gram without going significantly over. If it does go overweight, the soil must be dumped and weighing started over. Although the analytical sample doesn't need to be exactly 1 gram, it should be close. If a larger sized scoop was used and the amount of sample in the scoop looks larger than 1 gram. the technician may give the scoop a little shake to dump some of the larger particles back into the jar. This action selects AGAINST larger particles.

Now, say the 1st scoop of soil brought the balance to 0.7 g. Then a smaller volume (with even fewer large particles that might "tip the weight over") may be scooped into the weigh boat. Say that now the balance says 0.9 g. To get the mass closer to 1.0, the technician will likely gently tap the side of the scoop while holding it over the weight boat in order to knock smaller particles in a little at a time. This action selects AGAINST larger, low-Pb content particles and preferentially adds smaller, high Pb-content particles.

These very common techniques are fine when weighing out materials that are truly homogenous and have a uniform particle size. But for soils, variable selection for and against various particle sizes in the analyzed subsample changes the result. These various weighing techniques may all occur in the same sample weighing, or only one or none may occur. The fact that these variables are not controlled in routine laboratory practice is part of the reason why split sample results can be very different, and explains why lab duplicates from the same jar often have poor precision

Size conversions: 3/8" = 0.375 in. = 9.525 mm ASTM (US std) nominal aperture mesh size (mm): 4-mesh = 4.76 mm 10-mesh = 2 mm 50-mesh = 0.297 = 0.3 mm 200-mesh = 0.074 mm







The nugget effect can occur when contamination occurs in particulate form (such as explosives residues deposited as a powder or lead fragments in a firing range), or when contaminants partition onto mineral surfaces or organic carbon which are themselves heterogeneously distributed. Gy theory relates the size of the matrix particles to the mass of a sample support that can be representative of the true mean for the larger matrix volume.

Specifying a regulatory threshold without specifying the sample support over which it applies (or at least recognizing that differences in sample support introduce variability into analytical data results) easily leads to widely different analytical results. Since the sample support is generally ignored in regulation, it is ignored in practice and the sample support is left to chance. This leads to uncontrolled (and usually undocumented) variations in sampling conditions and often widely varying results that are difficult to interpret. Unless the lab was in charge of field sampling and was involved in project planning and SAP/QAPP preparation, the lab cannot be held accountable for such variable results. The analytical result is probably correct; project planning was faulty for not ensuring that sample collection procedures would produce samples representative of the decision.

The volume of the sample is an important factor that influences the reported concentration for the sample, especially when contaminants are heterogeneously distributed throughout the parent matrix. The nature of the release, such as contaminant release in the form of a powder, increases the heterogeneity. Alternatively, contaminants will be heterogeneously distributed if they preferentially partition onto mineral surfaces or into organic carbon that are themselves typically heterogeneously distributed.

3 different color-coded sample volumes are illustrated here. From largest to smallest: green/brown, light blue, and red. The dark particles are "contaminated" particles in a matrix of "cleaner" particles. Variable capture of the "dirty" particles is illustrated for higher and lower contaminant concentrations and different sample volumes.







Gilbert, Richard O. and Pamela G. Doctor. 1985. Determining the Number and Size of Soil Aliquots for Assessing Particulate Contaminant Concentrations. Journal of Environmental Quality Vol 14, No 2, pp. 286-292.





































When we think about measuring the characteristics of a human population (e.g., average height), the basic sampling unit is well-defined and obvious: a person.

Within the world of soil samples, however, there is no basic sampling unit that has been defined by guidance and is universally used. The sampling unit (or sampling support) varies from project to project; different sampling units are often used at different times for the same project.

At the physical scale that most sampling work is done, the soils that one collects, prepares, and analyzes are aggregates or "composites" consisting of lots of tiny soil particles. This is true for the pinch of soil used for an extraction, the heap of soil that the original pinch was drawn from, the jar from which the heap of soil came, and the soil core that yielded jar of soil.

Incremental vs. Composite Sampling Basic Differences (1)

- Goal of incremental sampling: estimate the average concentration over some defined area/volume of soil
- This defined soil area/volume is the subject of a decision of e.g., risk/no risk or exceedance/no exceedance
 - Therefore, the defined area/volume is called a decision unit (DU)
 - Having 3 or more independent replicates allows calculation of an upper confidence limit (UCL) on the mean (i.e., a conservative estimate of the mean concentration)

2/16/2012

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63





Incremental vs. Composite Sampling Basic Differences (2)

- Common goal of composite sampling: gain
 information about contaminants' spatial distribution
 - Accurate or conservative estimate of the concentration mean within the given area not required
 - The "given area" is called a sampling unit (SU); typically smaller than the DU
 - Composite replicates not typically used
 - Improves representativeness of "single" sample representing a small area by suppressing short-scale heterogeneity
- Composite uses fewer increments per sample than incremental sampling

2/16/2012

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66







G-5S p 121

In the first two cases (Sections 10.2 and 10.3), there is interest in *making an estimate for a prescribed target population*—in the first case estimating the mean of a continuous measure (e.g, the mean concentration of contaminant) and in the second case estimating the proportion of the population with a characteristic. In these two cases, carrying out the Composite sampling means combining a sampling design with a compositing protocol. The sampling design describes the method for selecting units from the target population and indicates the number of units to be selected and which ones are to be selected. The compositing protocol describes the scheme for forming and processing (mixing and homogenizing) composites. It indicates whether entire samples or aliquots are to be combined, the number of groups of units to be formed (m), the number of units per group (k), which units form each group, and the amount of material from each unit to be used in forming the composite sample.

The last two cases (Sections 10.4 and 10.5) involve *decision making at the unit level* rather than at the target population level. As a consequence, these approaches involve composite sampling and retesting protocols that not only define how composites are to be formed but also define when and how subsequent testing is to be done to ultimately identify particular units. The retesting strategies for these cases are conditional on the results obtained for the composites. In order to retest individual samples, the identity and integrity of the individual samples must be maintained; this implies that aliquots from the individual samples, rather than the whole samples, must be combined in forming composites. Additional aliquots from the individual samples are then retested either singly or in other composites.

RCRA Waste Sampling Draft Technical Guidance: Planning, Implementation, and Assessment, EPA530-D-02-002, August 2002 (formerly SW-846 Chapter 9, until pulled out to expand sampling into its own document) This document and SW-846 Method 8330B contain the most implementation details.






In 1995 EPA was Confident Compositing Would Catch On

In light of recent developments, <u>composite sampling is increasingly becoming an acceptable practice for sampling soils</u>, biota, and bulk materials when the goal is estimation of some population value under restrictions of a desired standard error and/or limits on the cost of sample measurement. In response to an informal survey of various professionals, several favorable applications of composite sampling were received. They include:

- Establishing and verifying attainment of remedial cleanup standards in soils using sample compositing and bootstrapping techniques
- Use of compositing to obtain adequate support in geostatistical sampling
- Optimal compositing strategies for screening material for deleterious agents
- A soil sample design utilizing techniques of compositing, binary search, and confidence limits on proportions
 Description

2/16/2012

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3.1. Soil Sampling	
3.1.1. Characterization of Soil PCB Contamination at Gas Pipeline Compressor Stations	
As part of a recent settlement between the Pennsylvania Department of Environmental Resources and the Texas Eastern Pipeline Company, PCB- contaminated soils had to be characterized and remediated at 19 sites. Be- cause waste sources included indiscriminate dumping, disposal in trash pits, air emissions and even application as weed killer along fence lines, the result- ing spatial distribution of contaminated soil was very heterogeneous, with hot spot locations unknown. Therefore, the only way to reliably characterize these sites required a very large number of soil samples, around 12,000 to be more precise. With each sample analyzed for total PCB, the cost for site characterization alone was around \$33 million. Now to really appreciate the magnitude of the problem, one must realize this discussion only pertains to the Pennsylvania settlement. The problem extends along the whole pipeline from the Gulf Coast to New England. Results of a retrospective study (Gore, Patil, and Taillie, 1992; Patil, Gore and Sinha, 1994), using the actual site characterization data, evealed that composite sampling methods potentially could have substantially reduced the analytical costs. Three aspects of the data were evaluated: (i) estimation of the mean and variance of total PCB concentration as well as total PCB mass, (ii) classification of each individual (uncomposited) sample as above or below a specified critical level, and (iii) quantification of those individual samples with the highest PCB levels.	Simulations run from actual site data show compositing produces accurate estimates of the mean but with much fewer analyses. pp. 8 & 9
Results showed that unbiased estimates of the mean and variance could be obtained with one fourth the number of analyses (90 instead of 360). A small loss of precision resulting from compositing seemed quite acceptable in light of large analytical cost reduction.	74

EPA Observational Economy Series: Composite Sampling (1995) pages 8 & 9

1995 EPA Compositing Guidance, p. 9

A site was acceptably clean if 90% of the measured samples were below 10 parts per million (ppm) with no values exceeding 25 ppm. With characterization data from the worst of the nineteen sites, compositing could have reduced the analytical cost of classifying individual samples according to the 10 ppm criterion by 9%, relative to exhaustive testing. Starting from this nearly worst case scenario, the cost savings increase as we move to cleaner sites and should be dramatic when analyzing post-remediation verification data. For example, another site along the pipeline that is cleaner, although still contaminated, could have had all individual samples classified according to 10 ppm for 50% less of the analytical cost associated with exhaustive testing. (See Gore, Boswell, Patil, and Taillie, 1992).

Finally, if concerned with simply knowing which individual sample has the highest concentration, we could have discovered this by exhaustively retesting just two composite samples. In other words, with only eight measurements in addition to the 90 composite measurements, we could have identified the "hottest" spot. Furthermore, 12 additional measurements could have revealed the locations with the four highest concentrations (See Patil, Gore and Sinha, 1994).

Instead of <u>360</u> analyses, 98 provides an unbiased estimate of mean & variance AND finds **the** highest hotspot

A total of 110 analyses would find mean + the **4** highest hotspots



EPA530-D-02-002 RCRA Waste Sampling Draft Technical

Guidance Planning, Implementation, and Assessment (Aug 2002) http://www.epa.gov/epawaste/hazard/testmethods/sw846/samp_guid.htm

ΟΡΤΙ	MIZING	THE DESIGN FOR OBTAINING THE DATA	50
5.1	Review	the Outputs of the First Six Steps of the DQO Process	50
5.2	Consid	er Data Collection Design Options	51
	5.2.1	Simple Random Sampling	57
	5.2.2	Stratified Random Sampling	57
	5.2.3	Systematic Sampling	-59
	5.2.4	Ranked Set Sampling	60
	5.2.5	Sequential Sampling	61
	5.2.6	Authoritative Sampling	62
		5.2.6.1 Judgmental Sampling	63
5.0	0	5262 Biased Sampling	64
5.3	Compo	Advantance and Limitations of Community Community	04
	5.3.1	Advantages and Limitations of Composite Sampling	00
	5.3.2 5.3.3	Composite Sampling Designs	67
		5.3.3.1 Simple Random Composite Sampling	67
		5.3.3.2 Systematic Composite Sampling	68
	5.3.4	Practical Considerations for Composite Sampling	- 69
	5.3.5	Using Composite Sampling To Obtain a More Precise Estimate of the	
		Mean	- 69
	5.3.6	Using Composite Sampling To Locate Extreme Values	
		or "Hot Spots"	71

	R	CRA Waste Sampling Draft Technical Guidance Planning, Implementation, and Assessment
6	CON	TROLLING VARIABILITY AND BIAS IN SAMPLING
	6.1 6.2 6.3	Sources of Random Variability and Bias in Sampling88Overview of Sampling Theory906.2.1 Heterogeneity906.2.2 Types of Sampling Error916.2.2.1 Fundamental Error926.2.2.2 Grouping and Segregation Error936.2.2.3 Increment Delimitation Error946.2.2.4 Increment Extraction Error946.2.2.5 Preparation Error946.2.3 The Concept of "Sample Support"94Practical Guidance for Reducing Sampling Error956.3.1 Determining the Optimal Mass of a Sample966.3.2 Obtaining the Correct Shape and Orientation of a Sample986.3.2.2 Sampling of a Moving Stream of Material996.3.3 Selecting Sampling Devices That Minimize Sampling Errors996.3.3 General Performance Goals for Sampling Tools and Devices99
	Pro	vides discussion of Gy theory and subsampling issues v 78



http://www.epa.gov/quality/qs-docs/g5s-final.pdf

10.	COM	POSITE SAMPLING
	10.1	OVERVIEW
	10.2	COMPOSITE SAMPLING FOR ESTIMATING A MEAN
		10.2.1 Overview
		10.2.2 Application 124
		10.2.3 Benefits "INCREMENTAL-AVERAGE" 125
		10.2.4 Limitations
		10.2.5 Implementation
		10.2.6 Relationship to Other Sampling Designs
		10.2.7 Examples
	10.3	COMPOSITE SAMPLING FOR ESTIMATING A POPULATION
		PROPORTION
		10.3.1 Overview
		10.3.2 Application
		10.3.3 Benefits
		10.3.4 Limitations
		10.3.5 Implementation 135
		10.3.6 Relationship to Other Sampling Designs
		10.3.7 Examples 137
	APPE	NDIX 10-A COST AND VARIANCE MODELS 138
	APPE	NDIX 10-B ESTIMATING A POPULATION PROPORTION 141
	the second se	

		Page
11. COI	MPOSITE SAMPLING FOR IDENTIFYING A TRAIT AND EXTREME	
SAM	IPLING UNITS	143
11.1	COMPOSITE SAMPLING FOR IDENTIFYING A TRAIT	143
	11.1.1 Overview	143
	11.1.2 Application	144
	11.1.3 Benefits	145
	11.1.4 Limitations	145
	11.1.5 Implementation.	145
	11.1.6 Relationship to Other Sampling Designs	149
	11.1.7 Examples	151
11.2	COMPOSITE SAMPLING AND RETESTING FOR IDENTIFYING	
	EXTREME SAMPLING UNITS	151
	11.2.1 Overview	151
	11.2.2 Application	153
	11.2.3 Benefits	153
	11.2.4 Limitations	153
	11.2.5 Implementation	153
	11.2.6 Relationship to Other Sampling Designs	154
	"COMPOSITE-SEARCHING"	
2/16/201	2 Clu-In Incremental-Composite Webinar Module 1.3	81



Here's the distinction between composite averaging and composite searching for the purposes of this presentation.

In the case of composite averaging (figure on the right), we collect multiple samples or soil increments from within a decision unit and then combine them into a composite sample for analysis. Our goal is to determine whether the average concentration within the decision unit is less than some cleanup criterion.

In the case of composite searching (figure on the left), we collect multiple samples or soil increments from either across decision units (as illustrated here) or within decision units and then combine them into a composite sample for analysis. Our goal is to determine whether any of those original increments might have had contamination levels above some specified threshold that would be indicative of the presence of contamination at levels of concern.

The balance of this discussion will focus on composite averaging. Composite searching will be discussed later (also referred to as adaptive compositing).

An important side point: "Dilution" is not a concern for composite averaging. "Dilution" is a concern for composite searching.





EPA/540/1-89/002 December 1989

Risk Assessment Guidance for Superfund Volume I Human Health Evaluation Manual (Part A)

Interim Final

RAGS Vol 1, Part A 1989

http://epa.gov/oswer/riskassessment/ragsa/index.htm

2/16/2012

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What Does RAGS 1989 Say?

Page 4-11

Heterogeneous nature of soils. One of the largest problems in sampling soil (or other solid materials) is that its generally heterogeneous nature makes collection of representative samples difficult (and compositing of samples virtually impossible -- see Section 4.6.3). Therefore, a large number of soil samples may be required to obtain sufficient data to calculate an exposure concentration. Composite samples sometimes are collected to obtain a more homogeneous sample of a particular area; however, as discussed in a later section, compositing samples also serves to mask contaminant hot spots (as well as areas of low contaminant concentration).

Designation of hot spots. Hot spots (i.e., areas of very high contaminant concentrations) may have a significant impact on direct contact exposures. The sampling plan should consider characterization of hot spots through extensive sampling, field screening, visual observations, or a combination of the above.

2/16/2012

Composite samples. Composite samples – sometimes referred to as continuous samples for air – combine subsamples from different locations and/or times. As such, composite samples may dilute or otherwise misrepresent concentrations at specific points and, therefore, should be avoided as the only inputs to a risk assessment. For media such as soil, sediment, and ground water, composite samples generally may be used to assess the presence or absence of contamination; however, they may be used in risk assessment only to represent average concentrations (and thus exposures) at a site. For example, "hot spots" cannot be determined using composite samples.

Page 4-19

This statement was written before other EPA guidance was issued that explained how to use compositing to find hotspots.

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United States Environmental Protection Agency Office of Solid Waste and Emergency Response Washington, D.C. 20460 Publication 9285.7-08l May 1992

Supplemental Guidance to RAGS: Calculating the Concentration Term

http://rais.ornl.gov/documents/UCLsEPASupGuidance.pdf

2/16/2012

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Calculating the Concentration Term – EPA 1992					
Page 3	Why use the UCL as the average concentration? Statistical confidence limits are the classical tool for addressing uncertainties of a distribution average. The 95 percent UCL of the arithmetic mean concentration is used as the average concentration, because it is not possible to know the true mean. The 95 percent UCL, therefore, accounts for uncertainties due to limited sampling data at Superfund sites. As sampling data become less limited at a site, uncertainties decrease, the UCL moves closer to the true mean, and exposure evaluations using either the mean or the UCL produce similar results. This concept is illustrated in Highlight 2 .				
2/16/2012	Clu-In Incremental-Composite Webinar Module 1.3 88				

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Calcula	ating the Concentration Term – EPA 1	992
	Page 3	
	Should a value other than the 95 percent UCL be used for the concentration?	
	A value other than the 95 percent UCL can be used, provided the risk assessor can document that high coverage of the true population mean occurs (i.e., the value equals or exceeds the true population mean with high probability). For exposure areas	
2/16/2012	Clu-In Incremental-Composite Webinar Module 1.3	89

Publication 9285.7-47 December 2001

Risk Assessment Guidance for Superfund: Volume I Human Health Evaluation Manual (Part D, Standardized Planning, Reporting, and Review of Superfund Risk Assessments)

Final

RAGS Vol 1, Part D 2001

http://epa.gov/oswer/riskassessment/ragsd/index.htm

2/16/2012

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RAGS 2001 guidance mentions compositing only once, but in a way that implies acceptance

Page 2-3

pathway (both current and future) and medium. The SAP should be accompanied by detailed sampling maps showing the location and type of samples (e.g., grab, composite, or duplicate). It is important to consider how sample results will be used to estimate exposure point concentrations. Background samples should be collected from appropriate areas (e.g., areas proximate to the site, free of potential contamination by site chemicals and similar to the site in topography, geology, meteorology, and other characteristics).

2/16/2012

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Calculating UCLs...Dec 2002 page 1 of a chemical in the environment. This concentration, commonly termed the exposure point concentration (EPC), is a conservative estimate of the average chemical concentration in an environmental medium. The EPC is determined for each individual exposure unit within a site. An exposure unit is the area throughout which a receptor moves and encounters an environmental medium for the duration of the exposure. Unless there is site-specific evidence to the contrary, an individual receptor is assumed to be equally exposed to media within all portions of the exposure unit over the time frame of the risk assessment. EPA recommends using the average concentration to represent "a reasonable estimate of the concentration likely to be contacted over time" (EPA 1989). The guidance previously issued by EPA in 1992, Supplemental Guidance to RAGS: Calculating the Concentration Term (EPA 1992), states that, "because of the uncertainty associated with estimating the true average concentration at a site, the 95 percent upper confidence limit (UCL) of the arithmetic mean should be used for this variable." Incremental sampling reduces uncertainty in estimates of the true average

2/16/2012

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SSG-TBD page 5

1.3.2 Exposure Assumptions. SSLs are risk-based concentrations derived from equations combining exposure assumptions with EPA toxicity data. The models and assumptions used to calculate SSLs were developed to be consistent with Superfund's concept of "reasonable maximum exposure" (RME) in the residential setting. The Superfund program's method to estimate the RME for chronic exposures on a site-specific basis is to combine an average exposure point concentration with reasonably conservative values for intake and duration in the exposure calculations (U.S. EPA,

SSG-TBD page 89

Note that the size, shape, and orientation of sampling volume (i.e., "support") for heterogenous media have a significant effect on reported measurement values. For instance, particle size has a varying affect on the transport and fate of contaminants in the environment and on the potential receptors. Because comparison of data from methods that are based on different supports can be difficult, defining the sampling support is important in the early stages of site characterization. This may be accomplished through the DQO process with existing knowledge of the site, contamination, and identification of the exposure pathways that need to be characterized. Refer to *Preparation of Soil Sampling Protocols: Sampling Techniques and Strategies* (U.S. EPA, 1992f) for more information about soil sampling support.

2/16/2012

Clu-In Incremental-Composite Webinar Module 1.3

SSG-TBD pages 89 & 90

Compositing. Because the objective of surface soil screening is to ensure that the mean contaminant concentration does not exceed the screening level, the physical "averaging" that occurs during compositing is consistent with the intended use of the data. Compositing allows a larger number of locations to be sampled while controlling analytical costs because several discrete samples are physically mixed (homogenized) and one or more subsamples are drawn from the mixture and submitted for analysis. If the individual samples in each composite are taken across the EA, each composite represents an estimate of the EA mean.

A practical constraint to compositing in some situations is the heterogeneity of the soil matrix. The efficiency and effectiveness of the mixing process may be hindered when soil particle sizes vary widely or when the soil matrix contains foreign objects, organic matter, viscous fluids, or sticky material. Soil samples should not be composited if matrix interference among contaminants is likely (e.g., when the presence of one contaminant biases analytical results for another).

Before individual specimens are composited for chemical analysis, the site manager should consider homogenizing and splitting each specimen. By compositing one portion of each specimen with the other specimens and storing one portion for potential future analysis, the spatial integrity of each specimen is maintained. If the concentration of a contaminant in a composite sample is high, the splits of the individual specimens from which it was composed can be analyzed discretely to determine which individual specimen(s) have high concentrations of the contaminant. This will permit the site manager to determine which portion within an EA is contaminated without making a repeat visit to the site.

2/16/2012

Clu-In Incremental-Composite Webinar Module 1.3



"Incremental Sampling" per ITRC
A structured sampling and subsampling protocol for representative & reproducible sampling of a well- defined soil area or volume (a decision unit) to obtain a concentration value representative of the average concentration over the entire decision unit
ITRC ISM-1 document at a URL to be announced; or do Internet search for "ITRC ISM"
2/16/2012 Clu-In Incremental-Composite Webinar Module 1.3 98

• ITRC = Interstate Technology and Regulatory Council (http://www.itrcweb.org)

• The ITRC ISM-1 document is a web-based document to accommodate a wealth of hyperlinks between sections and to additional information and popup definitions. The entire document will be able to be printed, but is over 600 pages long due to the number of appendices and case examples.





• In the case of ICS designs, collecting multiple increments of uniform mass/ volume in an unbiased manner within the decision unit involves 1 of 2 strategies: Collect an incremental sample across the entire decision unit so that a single incremental sample represents the entire decision unit, or collect a composite sample that represents the average concentration across a smaller area within the decision unit. For the latter case, the entire area of the DU is split between several areas called sampling units (SUs), each of which is represented by a composite sample. SUs preserve spatial information about contaminant concentrations, which can be accessed later should the need arise (such as the DU average exceeds a decision threshold such that cleanup is needed).

• Maintaining the chain of sample representativeness at each step of sample/ increment collection may require sieving, grinding and/or increasing the mass of the analytical subsample. This is as critical for discrete samples as for incremental-composite samples. If representativeness is lost at the subsampling stage, the work that went into ensuring representativeness as the field level is wasted.



• ASTM D6323 (2003) Standard Guide for Laboratory Subsampling of Media Related to Waste Management Activities (www.astm.org)

• The graph shows that as particle size gets smaller, the mass of the subsample needed to obtain good precision gets smaller. The reason why grinding to small particle sizes (like 75 microns) is important for metals is that very small analytical subsamples are used, usually between 0.5 to 2 grams of soil. Even if sieving to 2 mm is performed, when only 1 gram is used (blue curve), the <u>minimum</u> variability in the data (given as %RSD) is extreme. (recall that as %RSD increases, variability increases and precision decreases)

• As an example, say you have a sample with a true Pb concentration of 100 ppm. The soil was sieved through a 10-mesh sieve (2 mm screen size) and everything larger than 2 mm was removed from the sample. Great care was taken to subsample in a representative way, but only 1 gram of that sieved soil is taken for the analytical subsample. The BEST precision that can be expected (no matter how carefully the tech performs the sampling) for the 100 ppm sample is for results to fall within the range of 12 to 188 ppm 95% of the time. That is a very wide range.

• From this minimum starting point, the variability only gets worse if sloppy subsampling techniques are used.

• Increasing the analytical subsample mass to 10 grams (purple curve) significantly improves the minimum %RSD (i.e., the best possible precision) even if the largest particle size stays at 2 mm. If the true concentration of the sample is 100, and great care is exercised during subsampling, 95% of the time 10 gram subsamples will provide results in the range of 70 to 130 ppm.

• On the other hand, if the particle size is reduced to less than 1/10th of a mm, the minimum variability for a 1-gram subsample is around 5 %RSD. This means that results for a 100 ppm sample could fall, at best, between 90 and 110 ppm 95% of the time.



Off-the-shelf coffee grinders and mortar/pestles are examples of tools that can aid disaggregation.

Heavy-duty grinders/mills that can reduce particle size to <100 microns include: rotary pulverizers, ball mills, and puck mills. For more information on sample processing, see ITRC ISM-1 Section 6.2.2.5

Gy theory is the theory of sampling particulate materials. More information concerning Gy theory can be found in the ITRC ISM-1 document, EPA/600/ R-03/027 (subsampling guidance) and EPA/530-D-02-002 (SW-846/RCRA sampling guidance).





Ref: Guidance for Obtaining Representative Laboratory Analytical Subsamples from Particulate Laboratory Samples EPA/600/R-03/027 Nov 2003

• An incorrect subsampling tool allows some portions of the soil sample population to have a greater probability than others of inclusion in the analytical subsample. The soil sample population is the portion of the matrix that has been defined as the target, such as all particles smaller than 2 mm diameter. If all particles less than 2 mm is the target population, the subsampling tool cannot discriminate against (for example) particles in the 1-to 2-mm range vs. particles smaller than 1 mm.

• One example of discrimination can happen when the soil sample has segregated in the jar so that the larger particles are predominately at the top of the jar. Then when the soil in the jar is poured out into a "slab" for subsampling, the larger particles end up at the bottom of the slab. As depicted in the figure, the larger particles are less likely to be chosen by the rounded scoop.

• Another example of discrimination against larger particles can occur if the subsampling tool is shallow and narrow (as are some spatulas), so that larger particles roll off the tool.

• An example of discriminating against smaller particles can occur when soil is scooped off the top of jar contents which have segregated such that finer particles have mostly settled to the bottom out of reach of the sampling tool.



Guidance for Obtaining Representative Laboratory Analytical Subsamples from Particulate Laboratory Samples EPA/600/R-03/027 Nov 2003

• A tool is "correct" if, as the tool is pushed forward, particles are included or excluded with equal probability.





From EPA 600/R-03/027							
re re	From Table 8. Subsampling methods that are NOT recommended based on experimental evaluation						
Method	Typical Increment Size	Sensitivity to Grouping & Segregation	Moisture Content	Correct Sampling Possible	Agreement with Calculated s _{FE} ²	Comments	
Degenerate Fractional Shoveling	Medium to Large	Moderate to High	Dry to Moist	Yes, if Careful	Unlikely	Performance Tied to Lot Mass; Subject to Bias; N.R.	
Rolling and Quartering	Large	High	Dry	Yes, if Careful	Usually Not Close	Highly Variable; NR	
Coning and Quartering	Large	High	Dry	Yes, if Careful	Usually Not Close	Usually Biased; NR	
V-Blender	N.A.	High	Dry	N.A.	Very Unlikely	Problems with GE; N.R.	
Vibratory Spatula	Small	Very High	Dry	No	Not Close	Problems with GE; NR	
Grab Sampler	Variable	Very High	Dry to Moist	No	Not Close	Biased and Variable; N.R.	
2/16/2012 Clu-In Incremental-Composite Webinar Module 1.3					108		


