



Environmental Technology Verification Report

Immunoassay Kit

Strategic Diagnostics Inc. EnviroGard PCB Test Kit



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EPA Report No. 20874
August 1998



Superfund Innovative Technology
Evaluation Program



Notice

The U.S. Environmental Protection Agency (EPA), through its Office of Research and Development (ORD), and the U.S. Department of Energy's Environmental Management (EM) Program, funded and managed, through Interagency Agreement No. DW89937854 with Oak Ridge National Laboratory, the verification effort described herein. This report has been peer and administratively reviewed and has been approved for publication as an EPA document. Mention of trade names or commercial products does not constitute endorsement or recommendation for use of a specific product.



UNITED STATES ENVIRONMENTAL PROTECTION AGENCY

Office of Research and Development
Washington, D.C. 20460



**ENVIRONMENTAL TECHNOLOGY VERIFICATION PROGRAM
VERIFICATION STATEMENT**

TECHNOLOGY TYPE:	POLYCHLORINATED BIPHENYL (PCB) FIELD ANALYTICAL TECHNIQUES
APPLICATION:	MEASUREMENT OF PCBs IN SOILS AND SOLVENT EXTRACTS
TECHNOLOGY NAME:	EnviroGard PCB TEST KIT
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The U.S. Environmental Protection Agency (EPA) has created a program to facilitate the deployment of innovative technologies through performance verification and information dissemination. The goal of the Environmental Technology Verification (ETV) Program is to further environmental protection by substantially accelerating the acceptance and use of improved and more cost-effective technologies. The ETV Program is intended to assist and inform those involved in the design, distribution, permitting, and purchase of environmental technologies. This document summarizes the results of a demonstration of the Strategic Diagnostics Inc. (SDI) EnviroGard PCB test kit.

PROGRAM OPERATION

EPA, in partnership with recognized testing organizations, objectively and systematically evaluates the performance of innovative technologies. Together, with the full participation of the technology developer, they develop plans, conduct tests, collect and analyze data, and report findings. The evaluations are conducted according to a rigorous demonstration plan and established protocols for quality assurance. EPA's National Exposure Research Laboratory, which conducts demonstrations of field characterization and monitoring technologies, with the support of the U.S. Department of Energy's (DOE) Environmental Management (EM) program, selected Oak Ridge National Laboratory (ORNL) as the testing organization for the performance verification of polychlorinated biphenyls (PCBs) field analytical techniques.

DEMONSTRATION DESCRIPTION

In July 1997, the performance of six PCB field analytical techniques was determined under field conditions. Each technology was independently evaluated by comparing field analysis results with those obtained using approved reference methods. Performance evaluation (PE) samples were also used to assess independently the accuracy and comparability of each technology.

The demonstration was designed to detect and measure PCBs in soil and solvent extracts. The demonstration was conducted at ORNL in Oak Ridge, Tennessee, from July 22 through July 29, 1997. The study was conducted under two environmental conditions. The first site was outdoors, with naturally fluctuating temperatures and relative humidity conditions. The second site was inside a controlled environmental chamber, with generally cooler temperatures and lower relative humidities. Multiple soil types, collected from sites in Ohio, Kentucky, and Tennessee, were analyzed in this study. Solutions of PCBs were also analyzed to simulate extracted surface wipe samples. The results of the soil and extract analyses conducted under field conditions by the technology were compared with results from analyses of homogenous replicate samples conducted by conventional EPA SW-846 methodology in an approved reference laboratory. Details of the demonstration, including a data summary and discussion of results, may be found in the report

TECHNOLOGY DESCRIPTION

The EnviroGard PCB test kit is a competitive binding enzyme immunoassay that performs rapid, interval testing for PCBs in soils and solutions at specified action levels of 1, 5, 10, and 50 parts per million (ppm). These results are reported in intervals (e.g., < 1 ppm, 1 to 5 ppm, etc.) rather than in specific quantities (e.g., 6.7 ppm). The test kit is standardized using Aroclor 1248, but it can also detect Aroclors 1016, 1242, 1254, and 1260. The following items are needed to run a test: the EnviroGard PCB test kit, the Soil Extraction Bottle kit, the equipment contained in the Soil Field Lab, methanol, and water. The test procedure entails collecting a 5-g soil sample and extracting the PCBs from it using methanol. To initiate the PCB test, PCB-enzyme conjugate is added to the antibody-coated test tubes. The soil extract sample is then added to the test tube. After a 15-min incubation period, the tubes are rinsed and a color developing solution is added. Color development is inversely related to the PCB concentration (e.g. the darker the color, the less analyte PCB is present in the sample). PCBs are detected using a photometer that measures the absorbance of each tube.

VERIFICATION OF PERFORMANCE

The following performance characteristics of the EnviroGard PCB test kit were observed:

Throughput: Throughput was 18 samples/hour under outdoor conditions and 9 to 10 samples/hour under chamber conditions. This rate included sample preparation and analysis.

Ease of Use: Three operators analyzed samples during the demonstration, but the technology can be run by a single trained operator. Minimal training (2 to 4 h) is required to operate the EnviroGard kit, provided the user has a fundamental understanding of basic chemical and field analytical techniques.

Completeness: The EnviroGard kit generated results for all 232 PCB samples for a completeness of 100%.

Blank results: All of the blank soil samples were reported as the lowest reporting interval, which included zero; therefore, the percentage of false positive results was 0%. One false positive result (13%) was reported for the extract samples. The EnviroGard kit reported no false negative results.

Precision: The overall precision—based on the percentage of combined sample sets where all four replicates were reported as the same interval—was 38% for the PE soils, 47% for the environmental soils, and 67% for the extracts.

Accuracy: Accuracy was assessed using PE soil and extract samples. Accuracy, defined as the percentage of EnviroGard results that agreed with the accepted concentration, was 51% for PE soils and 58% for extracts. In general, the percentage of samples that was biased high was much greater (47% for PE soils and 38% for extracts) than the percentage biased low (1% for PE soils and 4% for extracts).

Comparability: Comparability, like accuracy, was defined as the percentage of results that agreed with, was above (i.e., biased high), or was below (i.e., biased low) the reference laboratory result. The percentage of samples that agreed with the reference laboratory results was 53% for all soils (PE and environmental) and 63% for extracts. The percentage of samples that was biased high was again much greater (45% for soils and 38% for extracts) than the percentage that was biased low (2% for soils and 0% for extracts).

Regulatory Decision-making: One objective of this demonstration was to assess the technology's ability to perform at regulatory decision-making levels for PCBs, specifically 50 ppm for soils and 100 µg/100cm² for surface wipes. For PE and environmental soil samples in the range of 40 to 60 ppm, 39% of the EnviroGard results agreed with the reference laboratory. In contrast, 59% were biased high and 2% were biased low. For extract samples representing surface wipe sample concentrations of 100 µg/100 cm² and 1000 µg/100 cm² (assuming a 100cm wipe sample), 63% of the

EnviroGard results agreed with the extract spike concentration. In comparison, the percentage of extract samples that was biased high was 38%, and the percentage of samples that was biased low was 4%.

Data quality levels: The performance of the EnviroGard PCB test kit was characterized as biased and imprecise about 50% of the time, because nearly half of the data were biased relative to the accepted concentration values (in terms of accuracy) and had replicate results that were not reported as the same interval (in terms of precision). It should be noted that there was an increased likelihood of results being biased high as a result of the conservatism that the manufacturer has incorporated into the calculation of results.

The results of the demonstration show that the EnviroGard PCB test kit can provide useful, cost-effective data for environmental problem-solving and decision-making. Undoubtedly, it will be employed in a variety of applications, ranging from serving as a complement to data generated in a fixed analytical laboratory to generating data that will stand alone in the decision-making process. As with any technology selection, the user must determine if this technology is appropriate for the application and the project data quality objectives. For more information on this and other verified technologies, visit the ETV web site at <http://www.epa.gov/etv>.

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NOTICE: EPA verifications are based on an evaluation of technology performance under specific, predetermined criteria and the appropriate quality assurance procedures. EPA makes no expressed or implied warranties as to the performance of the technology and does not certify that a technology will always, under circumstances other than those tested, operate at the levels verified. The end user is solely responsible for complying with any and all applicable Federal, State, and Local requirements.



Foreword

The U.S. Environmental Protection Agency (EPA) is charged by Congress with protecting the nation's natural resources. The National Exposure Research Laboratory (NERL) is EPA's center for the investigation of technical and management approaches for identifying and quantifying risks to human health and the environment. NERL's research goals are to (1) develop and evaluate technologies for the characterization and monitoring of air, soil, and water; (2) support regulatory and policy decisions; and (3) provide the science support needed to ensure effective implementation of environmental regulations and strategies.

EPA created the Environmental Technology Verification (ETV) Program to facilitate the deployment of innovative technologies through performance verification and information dissemination. The goal of the ETV Program is to further environmental protection by substantially accelerating the acceptance and use of improved and cost-effective technologies. The ETV Program is intended to assist and inform those involved in the design, distribution, permitting, and purchase of environmental technologies. This program is administered by NERL's Environmental Sciences Division in Las Vegas, Nevada.

The U.S. Department of Energy's (DOE's) Environmental Management (EM) program has entered into active partnership with EPA, providing cooperative technical management and funding support. DOE EM realizes that its goals for rapid and cost-effective cleanup hinge on the deployment of innovative environmental characterization and monitoring technologies. To this end, DOE EM shares the goals and objectives of the ETV.

Candidate technologies for these programs originate from the private sector and must be commercially ready. Through the ETV Program, developers are given the opportunity to conduct rigorous demonstrations of their technologies under realistic field conditions. By completing the evaluation and distributing the results, EPA establishes a baseline for acceptance and use of these technologies.

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Abstract

In July 1997, the U.S. Environmental Protection Agency (EPA) conducted a demonstration of polychlorinated biphenyl (PCB) field analytical techniques. The purpose of this demonstration was to evaluate field analytical technologies capable of detecting and quantifying PCBs in soils and solvent extracts. The fundamental objectives of this demonstration were (1) to obtain technology performance information using environmental and quality control samples, (2) to determine how comparable the developer field analytical results were with conventional reference laboratory results, and (3) to report on the logistical operation of the technology. The demonstration design was subjected to extensive review and comment by EPA's National Exposure Research Laboratory (NERL) Environmental Sciences Division in Las Vegas, Nevada; Oak Ridge National Laboratory (ORNL); EPA Regional Offices; the U.S. Department of Energy (DOE); and the technology developers.

The demonstration study was conducted at ORNL under two sets of environmental conditions. The first site was outdoors, with naturally variable temperature and relative humidity conditions typical of eastern Tennessee in the summer. A second site was located inside a controlled environmental chamber having lower, and relatively stable, temperature and relative humidity conditions. The test samples analyzed during this demonstration were performance evaluation (PE) soil, environmental soil, and extract samples. Actual environmental soil samples, collected from sites in Ohio, Kentucky, and Tennessee, were analyzed and ranged in concentration from 0.1 to 700 parts per million (ppm). Extract samples were used to simulate surface wipe samples, and were evaluated at concentrations ranging from 0 to 100 µg/mL. The reference laboratory method used to evaluate the comparability of data was EPA SW-846 Method 8081.

The field analytical technologies tested in this demonstration were the L2000 PCB/Chloride Analyzer (Dexsil Corporation), the PCB Immunoassay Kit (Hach Company), the 4100 Vapor Detector (Electronic Sensor Technology), and three immunoassay kits: D TECH, EnviroGard, and RaPID Assay System (Strategic Diagnostics Inc.). The purpose of an Environmental Technology Verification Report (ETVR) is to document the demonstration activities, present demonstration data, and verify the performance of the technology. This ETVR presents information regarding the performance of SDI's EnviroGard PCB test kit. Separate ETVRs have been published for the other technologies demonstrated.

The EnviroGard PCB test kit is an immunoassay kit used to determine PCB concentrations as interval threshold values. The EnviroGard PCB test kit uses a competitive binding enzyme immunoassay to perform rapid, interval testing for PCBs in soils and solutions at specified action levels of 1, 5, 10, and 50 ppm. The test kit is standardized using Aroclor 1248, but it can also detect Aroclors 1016, 1242, 1254, and 1260. The presence of PCBs is detected by a photometer based on a colored reaction in which the color development is inversely proportional to the concentration of PCB in the sample (e.g., the darker the color, the less analyte PCB is present in the sample). The EnviroGard kit provides no information on Aroclor identification.

The EnviroGard's quantitative results were based on the analysis of threshold standards with every batch of 12 samples. Because the EnviroGard kit was an interval technique, method detection limits are not applicable. Precision, defined as the percentage of the sample sets where all four replicates were reported as the same interval range, was 38% for PE soils, 47% for environmental soils, and 67% for extracts. Accuracy, defined as the percentage of EnviroGard results that agreed with the accepted concentration, was 51% for PE soils and 58% for extracts. In general, the percentage of results that was biased high was much greater (47% for PE soils and 38% for extracts) than the percentage of samples that was biased low (1% for PE soils and 4% for extracts). Comparability was defined similarly to accuracy, but the EnviroGard result was compared with the reference laboratory result rather than with the accepted concentration to determine comparability. For all soil samples (PE and environmental), the percentage of EnviroGard results that agreed with the reference laboratory results was 53%, and the percentage that was biased high was much greater than the percentage biased low.

The demonstration found that the EnviroGard kit was simple to operate in the field, requiring about an hour for initial set-up and preparation for sample analysis. Once the kit was operational, the sample throughput of the EnviroGard kit was 18 samples/hour under outdoor conditions and 9 to 10 samples/ hour under chamber conditions. Three operators analyzed samples during the demonstration, but the technology can be run by a single, trained operator. Minimal training (2 to 4 h) is required to operate the EnviroGard kit, provided the user has a fundamental understanding of basic chemical and field analytical techniques. The overall performance of the EnviroGard PCB test kit was characterized as biased and imprecise about 50% of the time; however, the kit generated no false positive or false negative results for soil samples. It should be noted that there was an increased likelihood that results would be biased high as a result of the conservatism that the manufacturer has incorporated into the calculation of results.

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List of Abbreviations and Acronyms

AL	action level
ANOVA	analysis of variance
ASTM	American Society for Testing and Materials
BHC	benzenehexachloride
C	concentration at which the false positive error rate is specified
CASD	Chemical and Analytical Sciences Division (ORNL)
CCV	continuing calibration verification standard
CSCT	Consortium for Site Characterization Technology
DCB	decachlorobiphenyl
DOE	U. S. Department of Energy
DQO	data quality objective
ELISA	enzyme-linked immunosorbent assay
EM	Environmental Management (DOE)
EPA	U. S. Environmental Protection Agency
ERA	Environmental Resource Associates
EST	Electronic Sensor Technology
ETTP	East Tennessee Technology Park
ETV	Environmental Technology Verification Program
ETVR	Environmental Technology Verification Report
EvTEC	Environmental Technology Evaluation Center
fn	false negative result
FN	false negative decision error rate

fp	false positive result
FP	false positive decision error rate
HEPA	high-efficiency particulate air
ID	identifier
LCS	laboratory control sample
LMER	Lockheed Martin Energy Research
LMES	Lockheed Martin Energy Systems
LV	Las Vegas
MDL	method detection limit
MS	matrix spike
MSD	matrix spike duplicate
n	number of samples
NERL	National Exposure Research Laboratory (EPA)
NCEPI	National Center for Environmental Publications and Information
NRC	Nuclear Regulatory Commission
ORD	Office of Research and Development (EPA)
ORNL	Oak Ridge National Laboratory
ORO	Oak Ridge Operations (DOE)
PARCC	precision, accuracy, representativeness, completeness, comparability
PCB	polychlorinated biphenyl
PE	performance evaluation
ppb	parts per billion
ppm	parts per million; equivalent units: mg/kg for soils and $\mu\text{g/mL}$ for extracts
Pr	probability
QA	quality assurance
QC	quality control
R^2	coefficient of determination

RDL	reporting detection limit
RH	relative humidity
RFD	request for disposal
RPD	relative percent difference
RSD	relative standard deviation (percent)
RT	regulatory threshold
S^2	variance for the measurement
SARA	Superfund Amendments and Reauthorization Act of 1986
SD	standard deviation
SDI	Strategic Diagnostics Inc.
SITE	Superfund Innovative Technology Evaluation
SMO	sample management office
SOP	standard operating procedure
SSM	synthetic soil matrix
TCMX	tetrachloro-m-xylene
TSCA	Toxic Substance Control Act
Z_{1-p}	the $(1-p)^{\text{th}}$ percentile for the standard normal distribution
%D	percent difference

Acknowledgments

The authors wish to acknowledge the support of all those who helped plan and conduct the demonstration, analyze the data, and prepare this report. In particular, we recognize the technical expertise of Mitchell Erickson (Environmental Measurements Laboratory), Viorica Lopez-Avila (Midwest Research Institute), and Robert F. O'Brien (Pacific Northwest National Laboratory), who were peer reviewers of this report. For internal peer review, we thank Stacy Barshick (ORNL); for technical support during the demonstration, Todd Skeen and Ralph Ilgner (ORNL); for site safety and health support, Kim Thomas, Marilyn Hanner, and Fred Smith (ORNL); for administrative support, Betty Maestas and Linda Plemmons (ORNL); for sample collection support, Wade Hollinger, Charlotte Schaefer, and Arlin Yeager (LMES), and Mike Rudacille and W. T. Wright (EET Corporation); for preliminary soil characterization support, Frank Gardner, John Zutman, and Bob Schlosser (ORNL, Grand Junction, Colo.); for sample management support, Angie McGee, Suzanne Johnson, and Mary Lane Moore (LMES); for providing performance evaluations samples, Michael Wilson (EPA's Office of Solid Waste and Emergency Response's Analytical Operations and Data Quality Center); and for technical guidance and project management of the demonstration, David Carden, Marty Atkins, and Regina Chung (DOE's Oak Ridge Operations Office), David Bottrell (DOE, Headquarters), Deana Crumbling (EPA's Technology Innovation Office), and Stephen Billets, Gary Robertson, and Eric Koglin (EPA's National Exposure Research Laboratory, Las Vegas, Nevada). The authors also acknowledge the participation of Strategic Diagnostics Inc., in particular, Craig Kostyshyn, Tim Lawruk, Chris Jones, and Penny Kosinski, who performed the analyses during the demonstration.

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Section 1

Introduction

The performance evaluation of innovative and alternative environmental technologies is an integral part of the U.S. Environmental Protection Agency's (EPA's) mission. Early efforts focused on evaluating technologies that supported the implementation of the Clean Air and Clean Water Acts. In 1987, the Agency began to evaluate the cost and performance of remediation and monitoring technologies under the Superfund Innovative Technology Evaluation (SITE) program. This was in response to the mandate in the Superfund Amendments and Reauthorization Act (SARA) of 1986. In 1990, the U.S. Technology Policy was announced. This policy placed a renewed emphasis on "making the best use of technology in achieving the national goals of improved quality of life for all Americans, continued economic growth, and national security." In the spirit of the Technology Policy, the Agency began to direct a portion of its resources toward the promotion, recognition, acceptance, and use of U.S.-developed innovative environmental technologies both domestically and abroad.

The Environmental Technology Verification (ETV) Program was created by the Agency to facilitate the deployment of innovative technologies through performance verification and information dissemination. The goal of the ETV Program is to further environmental protection by substantially accelerating the acceptance and use of improved and cost-effective technologies. The ETV Program is intended to assist and inform those involved in the design, distribution, permitting, and purchase of environmental technologies. The ETV Program capitalizes upon and applies the lessons that were learned in the implementation of the SITE Program to the verification of twelve categories of environmental technology: Drinking Water Systems, Pollution Prevention/Waste Treatment, Pollution Prevention/ Innovative Coatings and Coatings Equipment, Indoor Air Products, Air Pollution Control, Advanced Monitoring Systems, EvTEC (an independent, private-sector approach), Wet Weather Flow Technologies, Pollution Prevention/Metal Finishing, Source Water Protection Technologies, Site Characterization and Monitoring Technology [also referred to as the Consortium for Site Characterization Technology (CSCT)], and Climate Change Technologies. The performance verification contained in this report was based on the data collected during a demonstration of polychlorinated biphenyl (PCB) field analytical technologies. The demonstration was administered by CSCT.

For each pilot, EPA utilizes the expertise of partner "verification organizations" to design efficient procedures for conducting performance tests of environmental technologies. To date, EPA has partnered with federal laboratories and state, university, and private sector entities. Verification organizations oversee and report verification activities based on testing and quality assurance protocols developed with input from all major stakeholder/customer groups associated with the technology area.

In July 1997, CSCT, in cooperation with the U.S. Department of Energy's (DOE's) Environmental Management (EM) Program, conducted a demonstration to verify the performance of six field analytical technologies for PCBs: the L2000 PCB/Chloride Analyzer (Dexsil Corporation), the PCB Immunoassay Kit (Hach Company), the 4100 Vapor Detector (Electronic Sensor Technology), and three immunoassay kits from Strategic Diagnostics Inc. (SDI): D TECH, EnviroGard, and RaPID Assay System. This environmental technology verification report (ETVR) presents the results of the demonstration study for one PCB field analytical technology, SDI's EnviroGard PCB test kit. Separate ETVRs have been published for the other five technologies.

Technology Verification Process

The technology verification process is intended to serve as a template for conducting technology demonstrations that will generate high-quality data that EPA can use to verify technology performance. Four key steps are inherent in the process:

- Needs identification and technology selection
- Demonstration planning and implementation
- Report preparation
- Information distribution

Needs Identification and Technology Selection

The first aspect of the technology verification process is to determine technology needs of EPA and the regulated community. EPA, DOE, the U.S. Department of Defense, industry, and state agencies are asked to identify technology needs and interest in a technology. Once a technology need is established, a search is conducted to identify suitable technologies that will address this need. The technology search and identification process consists of reviewing responses to *Commerce Business Daily* announcements, searches of industry and trade publications, attendance at related conferences, and leads from technology developers. Characterization and monitoring technologies are evaluated against the following criteria:

- meets user needs;
- may be used in the field or in a mobile laboratory;
- is applicable to a variety of environmentally impacted sites;
- has high potential for resolving problems for which current methods are unsatisfactory;
- is cost competitive with current methods;
- performs better than current methods in areas such as data quality, sample preparation, or analytical turnaround time;

-
- uses techniques that are easier and safer than current methods;
 - is a commercially available, field-ready technology.

Demonstration Planning and Implementation

After a technology has been selected, EPA, the verification organization, and the developer agree to the responsibilities for conducting the demonstration and evaluating the technology. The following tasks are undertaken at this time:

- identifying demonstration sites that will provide the appropriate physical or chemical environment, including contaminated media;
- identifying and defining the roles of demonstration participants, observers, and reviewers;
- determining logistical and support requirements (for example, field equipment, power and water sources, mobile laboratory, communications network);
- arranging analytical and sampling support;
- preparing and implementing a demonstration plan that addresses the experimental design, sampling design, quality assurance/quality control (QA/QC), health and safety considerations, scheduling of field and laboratory operations, data analysis procedures, and reporting requirements.

Report Preparation

Innovative technologies are evaluated independently and, when possible, against conventional technologies. The field technologies are operated by the developers in the presence of independent technology observers. The technology observers are provided by EPA or a third-party group. Demonstration data are used to evaluate the capabilities, limitations, and field applications of each technology. Following the demonstration, all raw and reduced data used to evaluate each technology are compiled into a technology evaluation report, which is mandated by EPA as a record of the demonstration. A data summary and detailed evaluation of each technology are published in an ETVR.

Information Distribution

The goal of the information distribution strategy is to ensure that ETVRs are readily available to interested parties through traditional data distribution pathways, such as printed documents. Documents are also available on the World Wide Web through the ETV Web site (<http://www.epa.gov/etv>) and through a Web site supported by the EPA Office of Solid Waste and Emergency Response's Technology Innovation Office (<http://CLU-in.com>).

Demonstration Purpose

The purpose of this demonstration was to obtain performance information for PCB field analytical technologies, to compare the results with conventional fixed-laboratory results, and to provide supplemental information (e.g., cost, sample throughput, and training requirements) regarding the operation of the technology. The demonstration was conducted under two climatic conditions. One set of activities was conducted outdoors, with naturally fluctuating temperatures and relative humidity conditions. A second set was conducted in a controlled environmental facility, with lower, relatively stable temperatures and relative humidities. Multiple soil types, collected from sites in Ohio, Kentucky, and Tennessee, were used in this study. PCB soil concentrations ranged from approximately 0.1 to 700 parts per million (ppm). Developers also analyzed 24 solutions of known PCB concentration that were used to simulate extracted wipe samples. The extract samples ranged in concentration from 0 to 100 µg/mL.

Section 2

Technology Description

Objective

The objective of this section is to describe the technology being demonstrated, including the operating principles underlying the technology and the overall approach to its use. The information provided here is excerpted from that provided by the developer. Performance characteristics described in this section are specified by the developer; they may or may not be substantiated by the data presented in Section 5.

Principle

The EnviroGard PCB test kit assay system for PCB analysis applies the principles of enzyme-linked immunosorbent assay (ELISA) to the determination of PCBs. In such an assay, an enzyme has been chemically linked to a PCB molecule or PCB analog to create a labeled PCB reagent. The labeled PCB reagent (called a conjugate) is mixed with an extract of native sample containing the PCB contaminant. A portion of the mixture is applied to a surface to which an antibody specific for PCB has been affixed. The native PCB and PCB-enzyme conjugate compete for a limited number of antibody sites. After a period of time, the solution is washed away, and what remains is either PCB-antibody complexes or enzyme-PCB-antibody complexes attached to the test surface. The proportion of the two complexes on the test surface is determined by the amount of native PCB in the original sample. The enzyme present on the test surface is used to catalyze a color change reaction in a solution added to the test surface. Because the amount of enzyme present is inversely proportional to the concentration of native PCB contaminant, the amount of color development is inversely proportional to the concentration of PCB contaminant.

The EnviroGard PCB test kit is a system that performs rapid, semi-quantitative testing for PCBs in soil and in solution at specified action levels of 1, 5, 10, and 50 ppm. The kit is standardized using Aroclor 1248, but this test can also detect Aroclors 1016, 1242, 1254, and 1260. The test screens for PCBs with 95% confidence of no false negatives at the action levels.

Method Overview

The following items are needed to run a test: the EnviroGard PCB test kit, the Soil Extraction Bottle kit, the equipment contained in the Soil Field Lab, methanol, and water. The test procedure entails collecting a 5-g soil sample and extracting the PCBs from it using methanol. To initiate the PCB test, PCB-enzyme conjugate is added to the antibody-coated test tubes. The soil extract is then added to the test tube. After a 15-min incubation period, the tubes are rinsed and a color developing solution is added. A photometer is used to measure the absorbance of each tube. Color development is inversely related to the PCB concentration (i.e., a darker color indicates less PCB than calibrator material).

Materials

The following is a list of the materials needed to perform an assay:

-
- EnviroGard PCB test kit (including 20 antibody-coated test tubes, four calibrators, and one negative control standard; store at 4 to 8 °C; do not freeze; check expiration date). Note: The stop solution in the kit is hydrochloric acid. Avoid contact with skin or eyes.
 - EnviroGard Soil Extraction Bottle kit (including weighing boats, wooden spatulas, soil extraction bottles containing three mixing beads, 20-cc syringe, syringe coupler, filter units, filter caps, 4-mL glass storage vials, stoppers, and blank labels). Note: This kit contains enough material for 14 soil samples.
 - EnviroGard Soil Field Lab (including photometer, charger, Repeater™ pipettes and tips, positive displacement pipette and tips, timer, wash bottle, test tube rack). Note: Make sure the portable balance works and has fresh batteries.
 - Portable balance
 - Methanol
 - Tap or distilled water
 - Marker with water-resistant ink
 - Lab coat, gloves, and goggles
 - Paper towels

Procedure

Extraction

Weigh out 5 g of soil using the portable balance, weigh boat, and wooden spatula. Uncap the soil extraction bottle and label it appropriately. Fold the weigh boat into the mouth of the bottle and pour in the soil sample. Attach the 50-mL tip to the Repeater pipette and set the dial to 5. (This is equivalent to 5 mL.) Add 5 mL of methanol to each bottle. (*Note: It may be necessary to add an additional 5 mL of methanol if the sample soaks up all of the methanol, leaving little or no excess liquid to decant.*) Screw the cap onto the extraction bottle and tighten. Agitate the bottle for 2 min. Remove and discard the extraction bottle cap. Tightly screw a filter cap on the bottle. Attach a filter unit to the filter cap. Draw air into the syringe by pulling the plunger to the 20 mL mark. Twist the syringe firmly onto the open end of the filter unit. Push down the plunger. This creates enough pressure in the soil extraction bottle to drive the soil extract through the filter. Hold the filter unit and twist off the syringe coupler to remove the syringe assembly. Immediately invert the pressurized extract bottle and insert the filter outlet into the mouth of the glass storage vial. Hold the vial steady (or place it in a rack) until the necessary amount of soil extract (a minimum of 25 µL) has been filtered. Cap the glass storage vial.

Assay

Remove up to 20 of the antibody-coated test tubes from the kit and label them according to Table 2-1.

Table 2-1. Test tube labels

Tube label	Indicates...
NC	negative control
1 ppm	1 ppm PCB calibrator ^a
5 ppm	5 ppm PCB calibrator ^a
10 ppm	10 ppm PCB calibrator ^a
50 ppm	50 ppm PCB calibrator ^a
S1	Sample 1
S2	Sample 2

^aThe calibrators have actual concentrations of 0.5, 3, 5, and 22 ppm of Aroclor 1248, respectively.

Using the Repeater pipette labeled “CON,” dispense 500 µL of enzyme conjugate into each tube. Using the positive displacement pipette, dispense 25 µL of the negative control, calibrators, and sample(s) into the appropriate test tubes. Shake the test tube rack thoroughly for 5 s to mix the contents. Then set the timer for 15 min and let the tubes incubate undisturbed.

After the incubation period, empty the test tube contents into a sink or suitable container. Fill the tubes with water. Then empty them and shake out the remaining drops. Repeat the wash three times. Then invert the tubes and tap them on paper towels to remove excess water. Using the Repeater pipette labeled “SUB,” dispense 500 µL of the substrate into each tube. Let the tubes incubate for 5 min. During this time, you should see the tubes turn varying shades of blue, depending on the PCB concentration. (*Note: If a blue color does not develop in the “NC” test tube within 5 min after adding the substrate, this test is invalid and must be repeated.*)

After 5 min, add 500 µL of the stop solution to each tube using the Repeater pipette labeled “STOP”. The tube contents should turn yellow. **Use caution when handling the stop solution because it is an acid.**

Add 1 mL of the stop solution or wash water to a new empty test tube . (This is the photometer “blank” tube.) Insert the blank tube into the left well of the photometer. Dry the outside of each labeled tube with a clean paper towel and place each tube (one by one) into the right well of the photometer. Record the absorbance (optical density) reading of each tube. The tube contents must be read with a photometer within 30 min after adding the stop solution. Dispose of the tube in an appropriate waste container.

Interpreting Results

Interpret the photometer readings using Table 2-2.

The EnviroGard PCB test kit is calibrated to Aroclor 1248. If it is known that other Aroclors are present in the sample, Table 2-3 should be used for result interpretation.

Table 2-2. Interpretation of photometer readings

A sample with an absorbance reading...	...contains...	...and is reported as...
> (the reading of the 1 ppm PCB calibrator)	< 1.0 ppm PCB	[0, 1)
≤ (the reading of the 1 ppm PCB calibrator) and ≥ (the reading of the 5 ppm PCB calibrator)	≥ 1 ppm PCB and ≤ 5 ppm PCB	[1, 5]
< (the reading of the 5 ppm PCB calibrator) and ≥ (the reading of the 10 ppm PCB calibrator)	> 5 ppm PCB and ≤ 10 ppm PCB	(5, 10]
< (the reading of the 10 ppm PCB calibrator) and ≥ (the reading of the 50 ppm PCB calibrator)	> 10 ppm PCB and ≤ 50 ppm PCB	(10, 50]
< (the reading of the 50 ppm PCB calibrator)	> 50 ppm PCB	(50, ∞)

Table 2-3. Converting to Aroclor-specific concentrations

Aroclor 1248	Aroclor 1260	Aroclor 1242	Aroclor 1254
1 ppm	2 ppm	2 ppm	1.1 ppm
5 ppm	10 ppm	10 ppm	5.5 ppm
10 ppm	20 ppm	20 ppm	11.1 ppm
50 ppm	100 ppm	100 ppm	55 ppm

Possible Interfering Compounds

The following substances were tested and found to have less than 0.5% weight-to-weight of the immunoreactivity of Aroclor 1248: 1,2-dichlorobenzene, 1,3-dichlorobenzene, 1,4-dichlorobenzene, 1,2,4-trichlorobenzene, 2,4-dichlorophenol, 2,5-dichlorophenol, 2,4,5-trichlorophenol, 2,4,6-trichlorophenol, biphenyl, and pentachlorophenol.

Section 3

Site Description and Demonstration Design

Objective

This section describes the demonstration site, the experimental design for the verification test, and the sampling plan (sample types analyzed and the collection and preparation strategies). Included in this section are the results from the predemonstration study and a description of the deviations made from the original demonstration design.

Demonstration Site Description

Site Name and Location

The demonstration of PCB field analytical technologies was conducted at Oak Ridge National Laboratory (ORNL) in Oak Ridge, Tennessee. PCB-contaminated soils from three DOE sites (Oak Ridge; Paducah, Kentucky; and Piketon, Ohio) were used in this demonstration. The soil samples used in this study were brought to the demonstration testing location for evaluation of the field analytical technologies.

Site History

Oak Ridge is located in the Tennessee River Valley, 25 miles northwest of Knoxville. Three DOE facilities are located in Oak Ridge: ORNL, the Oak Ridge Y-12 Plant, and East Tennessee Technology Park (ETTP). Chemical processing and warhead component production have occurred at the Y-12 Plant, and ETTP is a former gaseous diffusion uranium enrichment plant. At both facilities, industrial processing associated with nuclear weapons production has resulted in the production of millions of kilograms of PCB-contaminated soils. Two other DOE facilities—the Paducah plant in Paducah, Kentucky, and the Portsmouth plant in Piketon, Ohio—are also gaseous diffusion facilities with a history of PCB contamination. During the remediation of the PCB-contaminated areas at the three DOE sites, soils were excavated from the ground where the PCB contamination occurred, packaged in containers ranging in size from 55-gal to 110-gal drums, and stored as PCB waste. Samples from these repositories—referred to as “Oak Ridge,” “Portsmouth,” and “Paducah” samples in this report—were used in this demonstration.

In Oak Ridge, excavation activities occurred between 1991 and 1995. The Oak Ridge samples comprised PCB-contaminated soils from both Y-12 and ETTP. Five different sources of PCB contamination resulted in soil excavations from various dikes, drainage ditches, and catch basins. Some of the soils are EPA-listed hazardous waste due to the presence of other contaminants (e.g., diesel fuels).

A population of more than 5000 drums containing PCB-contaminated soils was generated from 1986 to 1987 during the remediation of the East Drainage Ditch at the Portsmouth Gaseous Diffusion Plant. The ditch was reported to have three primary sources of potential contamination: (1) treated effluent from a radioactive liquid treatment facility, (2) runoff from a biodegradation plot where waste oil and sludge were disposed of, and (3) storm sewer discharges. In addition, waste oil was reportedly used for weed control in the ditch. Aside from

PCB contamination, no other major hazardous contaminants were detected in these soils. Therefore, no EPA hazardous waste codes are assigned to this waste.

Twenty-nine drums of PCB-contaminated soils from the Paducah plant were generated as part of a spill cleanup activity at an organic waste storage area (C-746-R). The waste is considered a listed hazardous waste for spent solvents (EPA hazardous waste code F001) because it is known to contain trichloroethylene. Other volatile organic compounds, such as xylene, dichlorobenzene, and cresol, were also detected in the preliminary analyses of some of the Paducah samples.

Site Characteristics

PCB-contaminated environmental soil samples from Oak Ridge, Portsmouth, and Paducah were collected from waste containers at storage repositories at ETTP and Paducah. Many of the soils contained interfering compounds such as oils, fuels, and other chlorinated compounds (e.g., trichloroethylene). Specific sample descriptions of the environmental soils used in this demonstration are given in Appendix A. In addition, each sample was characterized in terms of its soil composition, pH, and total organic carbon content. Those results are summarized in Appendix B.

Field demonstration activities occurred at two sites at ORNL: a natural outdoor environment (the outdoor site) and inside a controlled environmental atmosphere chamber (the chamber site). Figure 3-1 shows a schematic map of a section of ORNL indicating the demonstration area where the outdoor field activities occurred. Generally, the average summer temperature in eastern Tennessee is 75.6°F, with July and August temperatures averaging 79.1°F and 76.8°F, respectively. Average temperatures during the testing periods ranged from 79 to 85°F, as shown in Appendix C. Studies were also conducted inside a controlled environmental atmosphere chamber, hereafter referred to as the “chamber,” located in Building 5507 at ORNL. Demonstration studies inside the chamber were used to evaluate performance under environmental conditions that were markedly different from the ambient outdoor conditions at the time of the test. Average temperatures in the chamber during the testing periods ranged from 55 to 70°F. The controlled experimental atmosphere facility consists of a room-size walk-in chamber 10 ft wide and 12 ft long with air processing equipment to control temperature and humidity. The chamber is equipped with an environmental control system, including reverse osmosis water purification that supplies the chamber humidity control system. High efficiency particulate air (HEPA) and activated charcoal filters are installed for recirculation and building exhaust filtration.

Experimental Design

The analytical challenge with PCB analysis is to quantify a complex mixture that may or may not resemble the original commercial product (i.e., Aroclor) because of environmental aging, and to report the result as a single number [1]. The primary objective of the verification test was to compare the performance of the field technology with laboratory-based measurements. Often, verification tests involve a direct one-to-one comparison of results from field-acquired samples. However, because sample heterogeneity can preclude replicate field or laboratory comparison, accuracy and precision data must often be derived from the analysis of QC and performance evaluation (PE) samples. In this study, replicates of all three sample types (QC, PE, and environmental soil) were analyzed. The ability to use environmental soils in the verification test was made possible because the samples, collected from drums

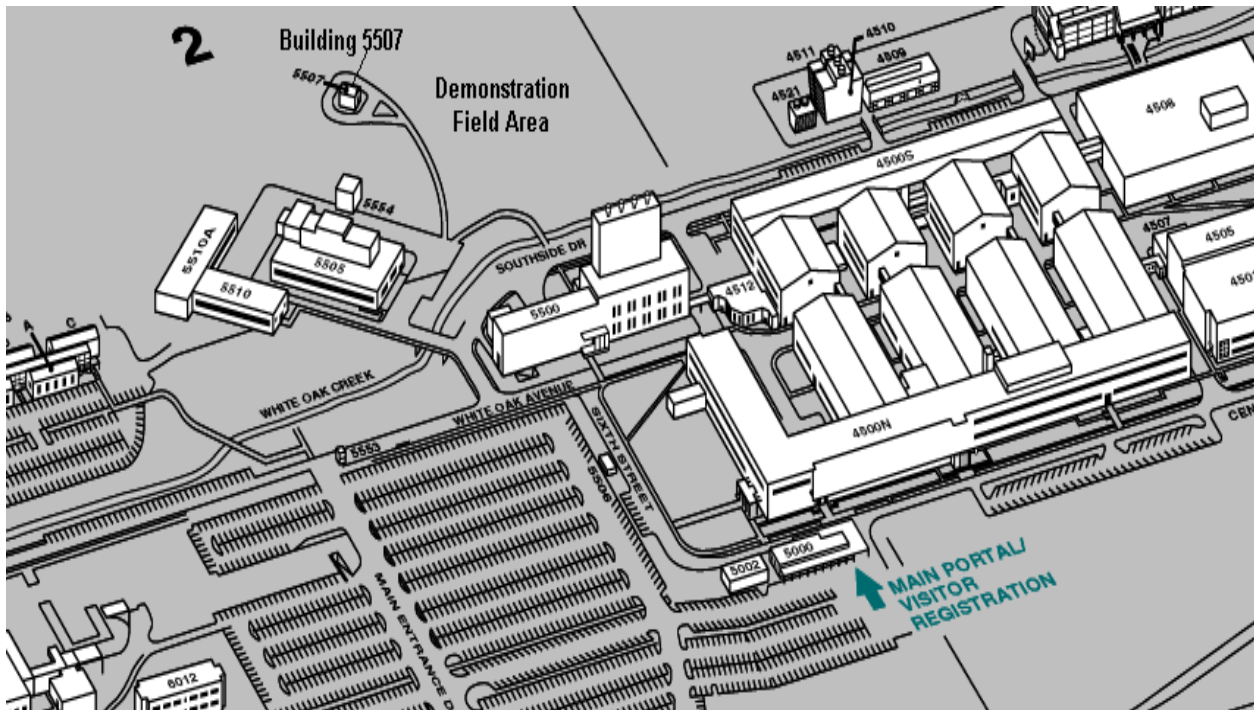


Figure 3-1. Schematic map of ORNL, indicating the demonstration area.

containing PCB-contaminated soils, could be thoroughly homogenized and characterized prior to the demonstration. This facet of the design, allowing additional precision data to be obtained on actual field-acquired samples, provided an added performance factor in the verification test.

Another objective of this demonstration was to evaluate the field technology’s capability to support regulatory compliance decisions. For field methods to be used in these decisions, the technology must be capable of informing the user, with known precision and accuracy, that soil concentrations are greater than or less than 50 ppm, and that wipe samples are greater than or less than 100 $\mu\text{g}/100 \text{ cm}^2$ [2]. The samples selected for analysis in the demonstration study were chosen with this objective in mind.

The experimental design is summarized in Table 3-1. This design was approved by all participants prior to the start of the demonstration study. In total, the developers analyzed 208 soil samples, 104 each at both locations (outdoors and chamber). The 104 soil samples comprised 68 environmental samples (17 unique environmental samples prepared in quadruplicate) ranging in PCB concentration from 0.1 to 700 ppm and 36 PE soils (9 unique PE samples in quadruplicate) ranging in PCB concentration from 0 to 50 ppm. To determine the impact of different environmental conditions on the technology’s performance, each batch of 104 samples contained 5 sets of quadruplicate soil samples from DOE’s Paducah site. These were analyzed under both sets of environmental conditions (i.e., outdoor and chamber conditions). For the developers participating in the extract sample portion (i.e., simulated wipe samples) of the demonstration, 12 extracts, ranging in concentration from 0 to 100 $\mu\text{g}/\text{mL}$, were analyzed in each location (chamber and outdoors). All samples were analyzed without prior knowledge of sample type or concentration and were analyzed in a randomized order that was unique for each developer.

Table 3-1. Summary of experimental design by sample type

Concentration range	Sample ID ^a		Total # samples analyzed
	Outdoor site	Chamber site	
<i>PE materials</i>			
0	126	226	8
2.0 ppm	118	218	8
2.0 ppm	124	224	8
5.0 ppm	120	220	8
10.9 ppm	122	222	8
20.0 ppm	119	219	8
49.8 ppm	125	225	8
50.0 ppm	121	221	8
50.0 ppm	123	223	8
<i>Environmental soils</i>			
0.1–2.0 ppm	101, 107, 108, 109, 113, 114	201, 202, 206	36
2.1–20.0 ppm	102, 103, 104, 115	203, 207, 212, 213	32
20.1–50.0 ppm	111, 116	204, 208, 209, 214, 215	28
50.1–700.0 ppm	105, 106, 110, 112, 117	205, 210, 211, 216, 217	40
<i>Extracts</i>			
0	129 ^b /132 ^c	229/232	8
10 µg/mL	127/130	227/230	8
100 µg/mL	128/131	228/231	8
Grand total	116	116	232 ^d

^a Each sample ID was analyzed in quadruplicate.

^b Extract prepared in iso-octane for Dextsil and the reference laboratory.

^c Extract prepared in methanol for Electronic Sensor Technology, Strategic Diagnostics Inc., and the reference laboratory.

^d All samples were analyzed in random order.

Environmental Conditions during Demonstration

As mentioned earlier, field activities were conducted both outdoors under natural environmental conditions and indoors in a controlled environmental atmosphere chamber to evaluate the effect of environmental conditions on technology performance. The weather outside was relatively uncomfortable during the July demonstration, with highs approaching 100°F and 90% relative humidity (RH). Daily average temperatures were around 85°F with 70% RH. While outside, the developers set up canopies to provide shade and protection from frequent late afternoon thundershowers.

In the indoor chamber tests, conditions were initially set to 55°F and 25% RH. An independent check of the conditions inside the chamber revealed that the temperature was closer to 68°F with a 38% RH on the first day of testing. A maintenance crew was called in to address the inconsistencies between the set and actual conditions. By the middle of the third day of testing, the chamber was operating properly at 55°F and 50% RH.

Appendix C contains a summary of the environmental conditions (temperature and relative humidity) during the demonstration. The SDI team analyzed samples using the EnviroGard test kit outdoors July 26 and 27 and in the chamber on July 23 and 24.

Sample Descriptions

PCBs ($C_{12}H_{10-x}Cl_x$) are a class of compounds that are chlorine-substituted linked benzene rings. There are 209 possible PCB compounds (also known as congeners). PCBs were commercially produced as complex mixtures beginning in 1929 for use in transformers, capacitors, paints, pesticides, and inks [1]. Monsanto Corporation marketed products that were mixtures of 20 to 60 PCB congeners under the trade name Aroclor. Aroclor mixtures are identified by a number (e.g., Aroclor 1260) that represents the mixture's chlorine composition as a percentage (e.g., 60%).

Performance Evaluation Materials

Samples of Tennessee reference soil [3] served as the blanks. Pre-prepared certified PE samples were obtained from Environmental Resource Associates (ERA) of Arvada, Colorado, and the Analytical Operations and Data Quality Center of EPA's Office of Solid Waste and Emergency Response. The soils purchased from ERA had been prepared using ERA's semivolatiles blank soil matrix. This matrix was a topsoil that had been dried, sieved, and homogenized. Particle size was approximately 60 mesh. The soil was approximately 40% clay. The samples acquired from EPA's Analytical Operations and Data Quality Center had been prepared using contaminated soils from various sites around the country in the following manner: The original soils had been homogenized and diluted with a synthetic soil matrix (SSM). The SSM had a known matrix of 6% gravel, 31% sand, and 43% silt/clay; the remaining 20% was topsoil. The dilution of the original soils was performed by mixing known amounts of contaminated soil with the SSM in a blender for no less than 12 h. The samples were also spiked with target pesticides (α , β , Δ , and δ -BHC, methoxychlor, and endrin ketone) to introduce some compounds that were likely to be present in an actual environmental soil. The hydrocarbon background from the original sample and the spiked pesticides produced a challenging matrix. The PE soils required no additional preparation by ORNL and were split for the developer and reference laboratory analyses as received.

Environmental Soil Samples

As noted in the site description, PCB-contaminated environmental soil samples from Oak Ridge, Portsmouth, and Paducah were used in this demonstration. The soils were contaminated with PCBs as the result of spills

and industrial processing activities at the various DOE facilities. Originally, the contaminated soils were excavated from dikes, drainage ditches, catch basins, and organic waste storage areas. The excavated soils were then packaged into waste containers and stored at the repositories in ETTP and Paducah in anticipation of disposal by incineration. The environmental soil samples used in this study were collected from these waste containers. Many of the soils contained interfering compounds such as oils, fuels, and other chlorinated compounds, while some contained multiple Aroclors. For more information on sampling locations and sample characteristics (soil composition, pH, and total organic carbon content), refer to Appendices A and B, respectively.

Extract Samples

Traditionally, the amount of PCBs on a contaminated surface is determined by wiping the surface with a cotton pad saturated with hexane. The pad is then taken to the laboratory, extracted with additional hexane, and analyzed by gas chromatography. Unlike soil samples, which can be more readily homogenized and divided, equivalent wipe samples (i.e., contaminated surfaces or post-wipe pads) were not easily obtainable. Therefore, interference-free solutions of PCBs were analyzed to simulate an extracted surface wipe pad. Extract sample analyses provided evaluation data that relied primarily on the technology's performance rather than on elements critical to the entire method (i.e., sample collection and preparation). Because different developers required the extract samples prepared in different solvents (e.g., methanol and iso-octane), the reference laboratory analyzed sets of extracts in both solvents. SDI analyzed extracts prepared in methanol. A total of 12 extracts were analyzed per site; these consisted of four replicates each of a blank and two concentration levels (10 and 100 µg/mL).

Sampling Plan

Sample Collection

Environmental soil samples were collected from April 17 through May 7, 1997. Portsmouth and Oak Ridge Reservation soils were collected from either storage boxes or 55-gal drums stored at ETTP. Briefly, the following procedure was used to collect the soil samples. Approximately 30 lb of soil were collected from the top of the drum or B-25 box using a scoop and placed in a plastic bag. The soil was sifted to remove rocks and other large debris, then poured into a plastic-lined 5-gal container. All samples were subjected to radiological screening and were determined to be nonradioactive. Similarly, soil samples were collected from 55-gal drums stored at Paducah and shipped to ORNL in lined 5-gal containers.

Sample Preparation, Labeling, and Distribution

Aliquots of several of the environmental soils were analyzed and determined to be heterogeneous in PCB concentration. Because this is unsatisfactory for accurately comparing the performance of the field technology with the laboratory-based method, the environmental soils had to be homogenized prior to sample distribution. Each Portsmouth and Oak Ridge environmental soil sample was homogenized by first placing approximately 1500 g of soil in a glass Pyrex dish. The dish was then placed in a large oven set at 35°C, with the exhaust and blower fans turned on to circulate the air. After drying overnight, the soil was pulverized using a conventional blender and sieved using a 9-mesh screen (2 mm particle size). Last, the soil was thoroughly mixed using a spatula. A comparison of dried and undried soils showed that a minimal amount of PCBs (< 20%) was lost as the result of sample drying, making this procedure suitable for use in the preparation of the soil samples. The Paducah samples, because of their sandy characteristics, required only the sieving and mixing preparation steps. Extract sample preparation involved making solutions of PCBs in methanol and iso-octane at two

concentration levels (10 and 100 µg/mL). Multiple aliquots of each sample were analyzed using the analytical procedure described below to confirm the homogeneity of the samples with respect to PCB concentration.

To provide the developers with soils contaminated at higher concentrations of PCBs, some of the environmental soils (those labeled with an “S” in Appendix B) were spiked with additional PCBs. Spiked soils samples were prepared after the soil was first dried in a 35°C oven overnight. The dry soil was ground using a conventional blender and sieved through a 9-mesh screen (2 mm particle size). Approximately 1500 g of the sieved soil were spiked with a diethyl ether solution of PCBs at the desired concentration. The fortified soil was agitated using a mechanical shaker and then allowed to air-dry in a laboratory hood overnight. A minimum of four aliquots were analyzed using the analytical procedure described below to confirm the homogeneity of the soil with regard to the PCB concentration.

The environmental soils were characterized at ORNL prior to the demonstration study. The procedure used to confirm the homogeneity of the soil samples entailed the extraction of 3 to 5 g of soil in a mixture of solvents (1 mL water, 4 mL methanol, and 5 mL hexane). After the soil/solvent mixture was agitated by a mechanical shaker, the hexane layer was removed and an aliquot was diluted for analysis. The hexane extract was analyzed on a Hewlett Packard 6890 gas chromatograph equipped with an electron capture detector and autosampler. The method used was a slightly modified version of EPA’s SW-846 dual-column Method 8081 [4].

After analysis confirming homogeneity, the samples were split into jars for distribution. Each 4-oz sample jar contained approximately 20 g of soil. Four replicate splits of each soil sample were prepared for each developer. The samples were randomized in two fashions. First, the order in which the filled jars were distributed was randomized so that the same developer did not always receive the first jar filled for a given sample set. Second, the order of analysis was randomized so that each developer analyzed the same set of samples, but in a different order. The extract samples were split into 10-mL aliquots and placed into 2-oz jars. The extracts were stored in the refrigerator (at ≤4°C) until released to the developers. Each sample jar had three labels: (1) developer order number, (2) sample identifier number, and (3) a PCB warning label. The developer order number corresponded to the order in which the developer was required to analyze the samples (e.g., SDI 1001 through SDI 1116). The sample identifier number was in the format of “xxx^yzz,” where “xxx” was the three-digit sample ID (e.g., 101) listed in Table 3-1, “y” was the replicate (e.g., 1 to 4), and “zz” was the aliquot order of each replicate (e.g., 01 to 11). For example, sample identifier 101101 corresponded to sample ID “101” (an Oak Ridge soil from RFD 40022, drum 02), “1” corresponded to the first replicate from that sample, and “01” corresponded to the first jar filled in that series.

Once the samples were prepared, they were stored at a central sample distribution center. During the demonstration study, developers were sent to the distribution center to pick up their samples. Samples were distributed sequentially in batches of 12 to ensure that samples were analyzed in the order specified. Completion of chain-of-custody forms and scanning of bar code labels documented sample transfer activities. Some of the developers received information regarding the samples prior to analysis. SDI received information pertaining to which Aroclors were in the samples and, if multiple Aroclors were present, in what ratio. This was provided at the request of SDI to simulate the type of information that would be available during actual field testing. The developers returned the unused portions of the samples with the analytical results to the distribution center when testing was completed. The sample bar codes were scanned upon return to document sample throughput time.

Three complete sets of extra samples, called archive samples, were available for distribution in case the integrity of a sample was compromised. Very few (<5) archive samples were utilized over the course of the demonstration.

Predemonstration Study

Ideally, environmental soil samples are sent to the developers prior to the demonstration study to allow them the opportunity to analyze representative samples in advance of the verification test. This gives developers the opportunity to refine and calibrate their technologies and revise their operating procedures on the basis of the predemonstration study results. The predemonstration study results can also be used as an indication that the selected technologies are of the appropriate level of maturity to participate in the demonstration study.

According to ORNL regulations, however, one of two conditions must exist in order to ship environmental soils that were once classified as mixed hazardous waste. First, the recipient—in this case, the developer’s facilities—must have proper Nuclear Regulatory Commission (NRC) licensing to receive and analyze radiological materials. Second, the soils must be certified as entirely free of radioactivity, beyond the no-rad certification issued from radiological screening tests based on ORNL standards. Because none of the developers had proper NRC licensing and proving that the soils were entirely free of radioactivity was prohibitive, spiked samples of Tennessee reference soil were used for the predemonstration study. The developers had an opportunity to evaluate the Tennessee reference soils spiked with PCBs at concentrations similar to what would be used in the demonstration study. The developers also analyzed two performance evaluation samples and one solvent extract. The reference laboratory analyzed the same set of samples, which included two extracts samples, prepared in the two solvents (methanol and iso-octane) requested by the developers.

Predemonstration Sample Preparation

Two soil samples were prepared by ORNL using Tennessee reference soil [3]. The soil was a Captina silt loam from Roane County, Tennessee, that was slightly acidic (pH ~5) and low in organic carbons (~1.5%). The soil composition was 7.7% sand, 29.8% clay, and 62.5% silt. To prepare a spiked sample, the soil was first ground using either a mortar and pestle or a conventional blender. The soil was then sieved through a 16-mesh screen (1 mm particle size). Approximately 500 g of the sieved soil was spiked with a diethyl ether solution of PCBs at the desired concentration. The soil was agitated using a mechanical shaker, then allowed to air-dry overnight in a laboratory hood. A minimum of five aliquots were analyzed by gas chromatography using electron capture detection. The PCB concentration of the spiked samples was determined to be homogeneous. The remaining two soil samples used in the predemonstration study were PE materials acquired from ERA and EPA (see the section “Performance Evaluation Materials” above). In addition, a solvent extract was prepared by ORNL to simulate an extracted surface wipe sample. The extracts were prepared in two different solvents (iso-octane and methanol) to accommodate developer requests.

Predemonstration Results

The predemonstration samples were sent to the developers and the reference laboratory on June 2, 1997. Predemonstration results were received by June 26, 1997. Table 3-2 summarizes the test kit’s results for the predemonstration samples. Results indicated that SDI’s EnviroGard PCB test kit was ready for field evaluation.

Table 3-2. Summary of EnviroGard predemonstration results

Sample description	Matrix	Source	EnviroGard ^a		Reference laboratory	
			Result (ppm)	Duplicate result (ppm)	Result (ppm)	Duplicate result (ppm)
2 ppm of Aroclor 1260	Soil	ORNL	[2, 8.2) ^b	[2, 8.2)	2.2	2.3
100 ppm (total) of Aroclors 1254 and 1260	Soil	ORNL	(85, ∞)	85, ∞)	78.0	89.0
11 ppm of Aroclor 1260	Soil	EPA	[8.2, 14.4)	[8.2, 14.4)	11.0	9.5
50 ppm of Aroclor 1254	Soil	ERA	(52, ∞) ^c		37.0 ^c	
5 ppm of Aroclor 1242	Extract	ORNL	(1.0, 5.0)	(1.0, 5.0)	4.7	4.9

^a Results were Aroclor-adjusted (see Section 2 for more details).

^b The notation [2, 8.2) indicates that the sample concentration was greater than or equal to 2 and less than 8.2. See Sections 2 and 5 for more information on interval reporting.

^c Replicate was not analyzed because of lack of adequate sample for second analyses.

Deviations from the Demonstration Plan

A few deviations from the demonstration plan occurred. In Appendix B of the technology demonstration plan [5], the reference laboratory’s procedure states that no more than 10 samples will be analyzed with each analytical batch (excluding blanks, standards, QC samples, and dilutions). The analytical batch is also stated as 10 samples in the Quality Assurance Project Plan of the demonstration plan. The reference laboratory actually analyzed 20 samples per analytical batch. Because a 20-sample batch is recommended in SW-846 Method 8081, this deviation was deemed acceptable.

Table 5 of the demonstration plan [5] delineates the environmental soils according to concentration. The classification was based on a preliminary analysis of the soils at ORNL. Table 3-1 of this report arranges the concentrations as characterized by the reference laboratory. The reference laboratory determined that five sample sets (sample IDs 102, 105, 110, 111, and 210) were in the next highest concentration range, differing from what was originally outlined in the demonstration plan. Also, the highest concentration determined by the reference laboratory was 700 ppm, while the preliminary analysis at ORNL found the highest concentration to be 500 ppm.

During the demonstration study, the SDI team did not note any deviations from the procedure described in the technology demonstration plan [5] for the EnviroGard test kit.

Section 4

Reference Laboratory Analytical Results and Evaluation

Objective and Approach

The purpose of this section is to present the evaluation of the PCB data generated by the reference laboratory. Evaluation of the results from the analysis of PE, environmental soil, and extract samples was based on precision, accuracy, representativeness, completeness, and comparability (PARCC) parameters [6]. This section describes how the analytical data generated by the reference laboratory were used to establish a baseline performance for PCB analysis.

Reference Laboratory Selection

The Oak Ridge Sample Management Office (SMO) has been tasked by DOE Oak Ridge Operations (DOE-ORO) with maintaining a list of qualified laboratories to provide analytical services. The technology demonstration plan [5] contains the SMO's standard operating procedures (SOPs) for identifying, qualifying, and selecting analytical laboratories. Laboratories are qualified as acceptable analytical service providers for the SMO by meeting specific requirements. These requirements include providing pertinent documentation (such as QA and chemical hygiene plans), acceptance of the documents by the SMO, and satisfactory performance on an on-site prequalification audit of laboratory operations. All laboratory qualifications are approved by a laboratory selection board, composed of the SMO operations manager and appointees from all prime contractors that conduct business with the SMO.

All of the qualified laboratories were invited to bid on the demonstration study sample analysis. The lowest-cost bidder was LAS Laboratories, in Las Vegas, Nevada. A readiness review conducted by ORNL and the SMO confirmed the selection of LAS as the reference laboratory. Acceptance of the reference laboratory was finalized by satisfactory performance in the predemonstration study (see Table 3-2). The SMO contracted LAS to provide full data packages for the demonstration study sample analyses within 30 days of sample shipment.

The SMO conducts on-site audits of LAS annually as part of the laboratory qualification program. At the time of selection, the most recent audit of LAS had occurred in February 1997. Results from this audit indicated that LAS was proficient in several areas, including program management, quality management, and training programs. No findings regarding PCB analytical procedure implementation were noted. A second on-site audit of LAS occurred August 11–12, 1997, during the analysis of the demonstration study samples. This surveillance focused specifically on the procedures that were currently in use for the analysis of the demonstration samples. The audit, jointly conducted by the SMO, DOE-ORO, and EPA-Las Vegas (LV), verified that LAS was procedurally compliant. The audit team noted that LAS had excellent adherence to the analytical protocols and that the staff were knowledgeable of the requirements of the method. No findings impacting data quality were noted in the audit report.

Reference Laboratory Method

The reference laboratory's analytical method, also presented in the technology demonstration plan [5], followed the guidelines established in EPA SW-846 Method 8081 [4]. According to LAS's SOP, PCBs were extracted from 30-g samples of soil by sonication in hexane. Each extract was then concentrated to a final volume that was further subjected to a sulfuric acid cleanup to remove potential interferences. The analytes were identified and quantified using a gas chromatograph equipped with dual electron-capture detectors. Each extract was analyzed on two different chromatographic columns with slightly different separation characteristics (primary column: RTX-1701, 30 m × 0.53 mm ID × 0.5 μm; confirmatory column: RTX-5, 30 m × 0.53 mm ID × 0.5 μm). PCBs were identified when peak patterns from a sample extract matched the patterns of standards for both columns. PCBs were quantified based on the initial calibration of the primary column.

Calibration

Method 8081 states that, because Aroclors 1016 and 1260 include many of the peaks represented in the other five Aroclor mixtures, it is only necessary to analyze two multilevel standards for these Aroclors to demonstrate the linearity of the detector response for PCBs. However, per LAS SOPs, five-point (0.1 to 4 ppm) initial calibration curves were generated for Aroclors 1016, 1248, 1254, and 1260 and the surrogate compounds [decachlorobiphenyl (DCB) and tetrachloro-m-xylene (TCMX)]. Single mid-level standards were analyzed for the other Aroclors (1221, 1232, and 1242) to aid in pattern recognition. All of the multi-point calibration data, fitted to quadratic models, met the QC requirement of having a coefficient of determination (R^2) of 0.99 or better over the calibration range specified. The detection limits for soil samples were 0.033 ppm (μg/g) for all Aroclors except Aroclor 1221, which was 0.067 ppm. For extract samples, the detection limits were 0.010 ppm (μg/mL) for all Aroclors except Aroclor 1221, which was 0.020 ppm. Reporting detection limits were calculated based on the above detection limits, the actual sample weight, and the dilution factor.

Sample Quantification

For sample quantification, Aroclors were identified by comparing the samples' peak patterns and retention times with those of the respective standards. Peak height ratios, peak shapes, sample weathering, and general similarity in detector response were also considered in the identification. Aroclor quantifications were performed by selecting three to five representative peaks, confirming that the peaks were within the established retention time windows, integrating the selected peaks, quantifying the peaks based on the calibrations, and averaging the results to obtain a single concentration value for the multicomponent Aroclor. If mixtures of Aroclors were suspected to be present, the sample was typically quantified in terms of the most representative Aroclor pattern. If the identification of multiple Aroclors was definitive, total PCBs in the sample were calculated by summing the concentrations of both Aroclors. Aroclor concentrations were quantified within the concentration range of the calibration curve. If PCBs were detected and the concentrations were outside of the calibration range, the sample was diluted and reanalyzed until the concentration was within the calibration range. If no PCBs were detected, the result was reported as a non-detect (i.e., "≤ reporting detection limit").

Sample Receipt, Handling, and Holding Times

The reference laboratory was scheduled to analyze a total of 256 PCB samples (208 soil samples, 24 iso-octane extract samples, and 24 methanol extract samples). Of these same samples, the developer was scheduled to analyze a total of 232 PCB samples (208 soil samples and 24 extract samples in solvent of choice). The samples were shipped to LAS at the start of the technology demonstration activities (July 22). Shipment was coordinated through the SMO. Completion of chain-of-custody forms documented sample transfer. The

amples were shipped on ice in coolers to maintain <6° mperatures during shipment. Samples were shipped with custody seals to ensure sample integrity and to prevent tampering during transport.

on receipt of the samples, the reference laboratory checked the receipt temperature and conditions of th sample containers, assigned each sample a unique number, an All samples were received at the proper temperature and in good condition. Demonstration samples wer divided into 11 analytical batches (with no order specified by ORNL to ensure that the analysis of sample types was ran supplied by the reference laboratory to indicate method performance, were performed with each analytica batch of soils.

Prior lysis, samples were stored in refrigerators kept at 4 to 6° e
r ct samples and to extract the soil samples within 14 days
o soils were extracted, the reference laboratory had an additional 40 days to
a r any of the demonstration samples. The
f es was received at ORNL in 72 days, on October 1, 1997.
The contractual obligation was 30 days.

he remainder of this section is devoted to summarizing the data generated by the reference laboratory and to

Quality Control Results

Objective

he purpose of this section is to provide an assessment of the data generated by the reference laboratory's QC rocedures. The QC samples included continuing calibration verification standards (CCVs), instrument blanks, ethod blanks, surrogate spikes, laboratory control samples (LCSs), and matrix spike (MS)/duplicate matrix pike (MSD) samples. Each control type is described in more detail in the following text and in the technology on plan [5]. Because extraction of these liquid samples was not required, calibration chec standards and instrument blanks were the only control samples implemented for the extract samples. Th reference laboratory's implementation of QC procedures was consistent with SW-846 guidance.

Continuing Calibration Verification Standard Results

CCV is a single calibration standard of known concentration, usually at the midpoint of the calibration range. s standard is evaluated as an unknown and is quantified against the initial calibration. The calculate concentration is then compared with th calibration is still valid. CCVs acceptance was QC requirement was for course of sample analysis.

Instrument and Method Blank Results

Instrument blanks (hexane) were analyzed prior to each CCV. The QC requirement was that instrument blanks must contain less than the reporting detection limit for any analyte. All instrument blanks were acceptable.

A method blank is an analyte-free soil matrix sample that is taken through the extraction process to verify that there are no laboratory sources of contamination. One method blank was analyzed for each analytical batch. The QC requirement was that method blanks must contain less than the reporting detection limit for any Aroclor. No PCBs were detected in any of the eleven method blanks that were analyzed. These results demonstrated that the reference laboratory was capable of maintaining sample integrity, and that it did not introduce PCB contamination to the samples during preparation.

Surrogate Spike Results

A surrogate is a compound that is chemically similar to the analyte group but is not expected to be present in the environmental sample. A surrogate is added to test the extraction and analysis methods to verify the ability to isolate, identify, and quantify a compound similar to the analyte(s) of interest without interfering with the determination. Two different surrogate compounds, DCB and TCMX, were used to bracket the retention time window anticipated in the Aroclor chromatograms. All soil samples, including QC samples, were spiked with surrogates at 0.030 ppm prior to extraction. Surrogate recoveries were deemed to be within QC requirements if the measured concentration fell within the QC acceptance limits that were established by past method performance. (For LAS this was 39 to 117% for DCB, and 66 to 128% for TCMX). The results were calculated using the following equation:

$$\text{percent recovery} = \frac{\text{measured amount}}{\text{actual amount}} \times 100\% \quad (4-1)$$

In all undiluted samples, both of the surrogates had percentage recoveries that were inside the acceptance limits. Surrogate recoveries in diluted samples were uninformative because the spike concentration (0.030 ppm, as specified by the method) was diluted below the instrument detection limits. The surrogate recovery results for undiluted samples indicated that there were no unusual matrix interferences or batch-processing errors for these samples.

Laboratory Control Sample Results

An LCS is an aliquot of a clean soil that is spiked with known quantities of target analytes. The LCS is spiked with the same analytes and at the same concentrations as the MS. (MSs are described in the next section.) If the results of the MS analyses are questionable (i.e., indicating a potential matrix effect), the LCS results are used to verify that the laboratory can perform the analysis in a clean, representative matrix.

Aroclors 1016 and 1260 were spiked into the clean soil matrix at approximately 0.300 ppm, according to the reference laboratory's SOP. The QC requirements (defined as percent recovery) for the LCS analyses were performance-based acceptance limits that ranged from 50 to 158%. In all but 1 of the 11 LCSs analyzed, both Aroclor percent recoveries fell within the acceptance limits. Satisfactory recoveries for LCS verified that the reference laboratory performed the analyses properly in a clean matrix.

Matrix Spike Results

In contrast to an LCS, a MS sample is an actual environmental soil sample into which target analytes are spiked at known concentrations. MS samples are used to assess the efficiency of the extraction and analytical methods for real samples. This is accomplished by determining the amount of spiked analyte that is quantitatively recovered from the environmental soil. An MSD sample is spiked and analyzed to provide a measure of method precision. Ideally, to evaluate the MS/MSD results, the environmental soil is analyzed unspiked so that the background concentrations of the analyte in the sample are considered in the recovery calculation.

For the demonstration study samples, one MS and MSD pair was analyzed with each analytical batch. The MS samples were spiked under the same conditions and QC requirements as the LCS (50 to 158% acceptance limits), so that MS/MSD and LCS results could be readily compared. The QC requirement for MS and MSD samples was a relative percent difference (RPD) of less than 30% between the MS/MSD pair. RPD is defined as

$$RPD = \frac{|MS\ recovery - MSD\ recovery|}{average\ recovery} \times 100\% \quad (4-2)$$

A total of 11 MS/MSD pairs were analyzed. Because the MS/MSD spiking technique was not always properly applied (e.g., a sample which contained 100 ppm of Aroclor 1254 was spiked ineffectively with 0.300 ppm of Aroclor 1260), many of the MS/MSD results were uninformative. For the samples that were spiked appropriately, all MS/MSD QC criteria were met.

Conclusions of the Quality Control Results

The reference laboratory results met performance acceptance requirements for all of the samples where proper QC procedures were implemented. Acceptable performance on QC samples indicated that the reference laboratory was capable of performing analyses properly.

Data Review and Validation

Objective

The purpose of validating the reference laboratory data was to ensure usability for the purposes of comparison with the demonstration technologies. The data generated by the reference laboratory were used as a baseline to assess the performance of the technologies for PCB analysis. The reference laboratory data were independently validated by ORNL and SMO personnel, who conducted a thorough quality check and reviewed all sample data for technical completeness and correctness.

Corrected Results

Approximately 8% of the results provided by the reference laboratory (20 of 256) were found to have correctable errors. So as not to bias the assessment of the technology's performance, errors in the reference laboratory data were corrected. These changes were made conservatively, based on the guidelines provided in the SW-846 Method 8081 for interpreting and calculating Aroclor results. The errors (see Appendix D, Table D-3) were categorized as transcription errors, calculation errors, and interpretation errors. The corrections

listed in Table D-3 were made in the final data set that was used for comparison with the demonstration technologies.

Suspect Results

Normally, one would not know if a single sample result was “suspect” unless (1) the sample was a PE sample, where the concentration is known or (2) a result was reported and flagged as suspect for some obvious reason (e.g., no quantitative result was determined). The experimental design implemented in this demonstration study provided an additional indication of the abnormality of data through the inspection of the replicate results from a homogenous soil sample set (i.e., four replicates were analyzed for each sample ID).

Data sets were considered suspect if the standard deviation (SD) of the four replicates was greater than 30 ppm and the percent relative standard deviation (RSD) was greater than 50%. Five data sets (sample IDs 106, 205, 216, 217, 225) contained measurements that were considered suspect using this criteria, and the suspect data are summarized in Table 4-1. A number of procedural errors may have caused the suspect measurements (e.g., spiking heterogeneity, extraction efficiencies, dilution, etc.). In the following subsections for precision and accuracy, the data were evaluated with and without these suspect values to represent the best and worst case scenarios.

Table 4-1. Suspect measurements within the reference laboratory data

Criteria	Sample ID	PCB concentration (ppm)		Data usability
		Replicate results (ppm)	Suspect result(s) (ppm)	
SD > 30 ppm and RSD > 50%	106	255.9, 269.9, 317.6	649.6	Performed data analysis with and without this value
	205	457.0, 483.3, 538.7	3,305.0	
	216	47.0, 54.3, 64.0	151.6	
	217	542.8, 549.8, 886.7	1,913.3	
	225	32.1, 36.5, 56.4	146.0	
Qualitative result	110	≤ reporting detection limits	≤ 66, ≤ 98, ≤ 99, ≤ 490	Used as special case for comparison with developer results
	112		≤ 66, ≤ 130, ≤ 200, ≤ 200	

Samples that did not fall into the above criteria, but were also considered suspect, were non-blank samples that could not be quantified and were reported as “≤ the reporting detection limit.” This was the case for environmental soil sample IDs 110 and 112. It is believed that the reference laboratory had trouble quantifying these soil samples because of the abundance of chemical interferences. These samples were diluted by orders of magnitude to reduce interferences, thereby diluting the PCB concentrations to levels that were lower than the instrument detection limits. With each dilution, the reporting detection limits values were adjusted for sample weight and dilution, which accounts for the higher reporting detection limits (up to 490 ppm). It is

believed that these samples should have been subjected to additional pre-analytical cleanup to remove these interferences before quantification was attempted. Sample IDs 110 and 112 were collected from the same cleanup site (see Appendix B), so it is not surprising that similar difficulties were encountered with both sample sets. Because the results for sample IDs 110 and 112 were not quantitative, these data were compared with the technology data only on a special case basis.

Data Assessment

Objective

The purpose of this section is to provide an evaluation of the performance of the reference laboratory results through statistical analysis of the data. The reference laboratory analyzed 72 PE, 136 environmental soil, and 48 extract samples. All reference laboratory analyses were performed under the same environmental conditions. Therefore, site differentiation was not a factor in data assessment for the reference laboratory. For comparison with the technology data, however, the reference laboratory data are delineated into “outdoor site” and “chamber site” in the following subsections. For consistency with the technology review, results from both sites were also combined to determine the reference laboratory’s overall performance for precision and accuracy. This performance assessment was based on the raw data compiled in Appendix D. All statistical tests were performed at a 5% significance level.

Precision

The term “precision” describes the reproducibility of measurements under a given set of conditions. The SD of four replicate PCB measurements was used to quantify the precision for each sample ID. SD is an absolute measurement of precision, regardless of the PCB concentration. To express the reproducibility relative to the average PCB concentration, RSD is used to quantify precision, according to the following equation:

$$RSD = \frac{\text{Standard Deviation}}{\text{Average Concentration}} \times 100\% \quad (4-3)$$

Performance Evaluation Samples

The PE samples were homogenous soils containing certified concentrations of PCBs. Results for these samples represent the best estimate of precision for soil samples analyzed in the demonstration study. Table 4-2 summarizes the precision of the reference laboratory for the analysis of PE samples. One suspect measurement (sample ID 225, 146.0 ppm) was reported for the PE soil samples. The RSDs for the combined data ranged from 9 to 33% when the suspect measurement was excluded, and from 9 to 79% including the suspect measurement. The overall precision, determined by the mean RSD for all PE

Table 4-2. Precision of the reference laboratory for PE soil samples

Outdoor Site				Chamber Site				Combined Sites		
Sample ID	Average concentration (ppm)	SD (ppm)	RSD (%)	Sample ID	Average concentration (ppm)	SD (ppm)	RSD (%)	Average concentration (ppm)	SD (ppm)	RSD (%)
126 ^a	0	n/a	n/a	226	0	n/a	n/a	0	n/a	n/a
118	1.6	0.6	39	218	2.6	0.2	6	2.1	0.7	33
124	1.7	0.2	13	224	1.7	0.5	29	1.7	0.4	21
120	5.0	1.0	20	220	5.8	1.8	31	5.4	1.4	26
122	11.1	0.9	8	222	12.8	0.3	3	11.9	1.1	9
119	20.1	3.4	17	219	23.3	6.1	26	21.7	4.9	23
125	37.9	6.9	18	225	41.7 ^b	12.9 ^b	31 ^b	39.5 ^c	9.2 ^c	23 ^c
121	54.6	3.4	6	221	44.9	11.3	25	49.8	9.3	19
123	60.1	4.6	8	223	55.8	7.7	14	58.0	6.3	11

^a All PCB concentrations were reported as non-detects.

^b Results excluding the suspect value (results including the suspect value: mean = 67.8 ppm, SD = 53.2 ppm, and RSD = 79%).

^c Results excluding the suspect value (results including the suspect value: mean = 52.8 ppm, SD = 38.6 ppm, and RSD = 73%).

samples, was 21% for the worst case (including the suspect result) and 18% for the best case (excluding the suspect result).

Environmental Soil Samples

The precision of the reference laboratory for the analysis of environmental soil samples is reported in Table 4-3. In this table, results including suspect measurements are presented in parentheses. Average concentrations were reported by the reference laboratory as ranging from 0.5 to 1,196 ppm with RSDs that ranged from 7 to 118% when the suspect results were included. Excluding the suspect results, the highest average concentration decreased to 660 ppm, and the largest RSD decreased to 71%. Because the majority of the samples fell below 125 ppm, precision was also assessed by partitioning the results into two ranges: low concentrations (< 125 ppm) and high concentrations (> 125 ppm). For the low concentrations, the average RSD was 23% excluding the suspect value and 26% including the suspect value. These average RSDs were only slightly larger than the RSDs for the PE soil samples of comparable concentration (18% for best case and 21% for worst case). Five soil sample sets (sample IDs: 106, 117, 205, 211 and 217) were in the high-concentration category. The average precision for high concentrations was 56% for the worst case and 19% for the best case. The precision estimates for the low and high concentration ranges were comparable when the suspect values were excluded. This indicated that the reference laboratory’s precision for the environmental soils was consistent (approximately 21% RSD) and was comparable to the PE soil samples when the suspect values were excluded.

The Paducah soils (indicated as bold sample IDs in Table 4-3) were analyzed by the technologies under both outdoor and chamber conditions to provide a measure of the effect that two different environmental

Table 4-3. Precision of the reference laboratory for environmental soil samples

Outdoor site				Chamber site			
Sample ID	Average concentration (ppm)	Standard deviation (ppm)	RSD (%)	Sample ID	Average concentration (ppm)	Standard deviation (ppm)	RSD (%)
101	0.5	0.1	16	206	1.9	0.9	49
102	2.0	0.3	16	207	18.8	3.5	19
103	2.3	0.6	27	208	30.5	7.9	26
104	9.4	4.0	43	209	40.2	28.5	71
105	59.4	16.5	28	210	88.6	25.6	29
106	281.0 (373.2) ^a	32.4 (186.2)	12 (50)	211	404.5	121.8	30
107	1.3	0.3	20	212	3.2	1.6	50
108	1.8	0.1	8	213	8.1	1.6	20
109	2.0	0.4	20	214	25.2	3.7	15
110	n/a ^b	n/a	n/a	215	26.7	3.2	12
111	38.7	4.3	11	216	55.1 (79.2)	8.5 (48.7)	15 (62)
112	n/a	n/a	n/a	217	659.8 (973.2)	196.6 (647.0)	30 (66)
113^c	1.1	0.6	55	201	0.9	0.2	24
114	1.3	0.3	20	202	1.4	0.2	12
115	14.8	1.8	12	203	13.9	1.7	12
116	41.3	5.9	14	204	44.3	2.9	7
117	383.9	55.2	14	205	493.0 (1196.0)	41.7 (1406.4)	8 (118)

^a Data in parentheses include suspect values.

^b n/a indicates that qualitative results only were reported for this sample.

^c Bold sample IDs were matching Paducah sample pairs (i.e., 113/201, 114/202, 115/203, 116/204, 117/205).

conditions had on the technology’s performance. Although this was not an issue for the reference laboratory (because all the samples were analyzed under laboratory conditions), the reference laboratory’s results were delineated into the different site categories for comparison with the technologies. Sample IDs 113 and 201, 114 and 202, 115 and 203, 116 and 204, and 117 and 205 each represent a set of eight replicate samples of the same Paducah soil. The RSDs for four of the five Paducah pairs (excluding the suspect value for sample ID 205) ranged from 11 to 17%. The result from one pair (sample IDs 113 and 201) had an RSD of 42%, but the reported average concentration was near the reporting limits.

Extract Samples

The extract samples, which were used to simulate surface wipe samples, were the simplest of all the demonstration samples to analyze because they required no extraction and were interference-free. Three types of extract samples were analyzed: solvent blanks, spikes of Aroclor 1242 at 10 µg/mL, and spikes of Aroclor 1254 at 100 µg/mL. Identical extract samples were prepared in two solvents (iso-octane and methanol) to accommodate the developer's request. The reference laboratory analyzed both solvent sets. A student's t-test [7, 8] was used to compare the reference laboratory's average PCB concentrations for the two different solvents and showed that no significant differences were observed at either concentration. Therefore, the reference laboratory results for the two extract solvents were combined. Additionally, all blank samples were quantified as non-detects by the reference laboratory.

Table 4-4 summarizes the reference laboratory results for the extract samples by site. RSDs for the four replicates for each sample ID ranged from 3 to 24%. For the combined data set (16 replicate measurements), the average RSD at the 10-µg/mL level was 19%, while the average RSD at the 100-µg/mL level was 8%. For the entire extract data set, an estimate of overall precision was 14%. The overall precision for the extract samples was comparable to the best-case precision for environmental soil samples (21%) and PE soil samples (18%).

Table 4-4. Precision of the reference laboratory for extract samples

Outdoor site				Chamber site				Combined sites		
Sample ID	Average conc (µg/mL)	SD (µg/mL)	RSD (%)	Sample ID	Average conc (µg/mL)	SD (µg/mL)	RSD (%)	Average conc (µg/mL)	SD (µg/mL)	RSD (%)
129 ^a	0	n/a	n/a	229	0	n/a	n/a	0	n/a	n/a
132 ^a	0	n/a	n/a	232	0	n/a	n/a			
127	10.9	0.4	4	227	9.6	0.8	8	10.4	1.9	19
130	12.1	2.9	24	230	8.9	1.4	16			
128	67.4	2.3	3	228	65.2	5.1	8	63.5	5.2	8
131	63.8	5.0	8	231	57.7	3.1	5			

^a All PCB concentrations reported as non-detects by the laboratory.

Accuracy

Accuracy represents the closeness of the reference laboratory's measured PCB concentrations to the accepted values. Accuracy was examined by comparing the measured PCB concentrations (for PE soil and extract samples) with the certified PE values and known spiked extract concentrations. Percent recovery was used to quantify the accuracy of the results. The optimum percent recovery value is 100%. Percent recovery values greater than 100% indicate results that are biased high, and values less than 100% indicate results that are biased low.

Performance Evaluation Soil Samples

The reference laboratory's performance for the PE samples is summarized in Table 4-5. Included in this table are the performance acceptance ranges and the certified PCB concentration values. The acceptance ranges, based on the analytical verification data, are guidelines established by the provider of the PE materials to gauge acceptable analytical results. As shown in Table 4-5, all of the average concentrations were within the acceptance ranges, with the exception of sample ID 218. The average result of sample ID 225 was outside of the acceptance range only when the suspect result was included. All of the replicate measurements in sample ID 225 were biased slightly high. Average percent recoveries for the PE samples (excluding suspect values) ranged from 76 to 130%. Overall accuracy was estimated as the average recovery for all PE samples. The overall percent recovery was 105% as a worst case when the suspect value was included. Excluding the suspect value as a best case slightly lowered the overall percent recovery to 101%. A regression analysis [9] indicated that the reference laboratory's results overall were unbiased estimates of the PE sample concentrations.

Table 4-5. Accuracy of the reference laboratory for PE soil samples

Certified concentration (ppm) (acceptance range, ppm)	Outdoor site			Chamber site			Combined sites	
	Sample ID	Average conc (ppm)	Recovery (%)	Sample ID	Average conc (ppm)	Recovery (%)	Average conc (ppm)	Recovery (%)
0 ^a (n/a)	126	0	n/a	226	0	n/a	0	n/a
2.0 (0.7-2.2)	118	1.6	79	218	2.6	130	2.1	105
2.0 (0.9-2.5)	124	1.7	85	224	1.7	85	1.7	85
5.0 (2.1-6.2)	120	5.0	99	220	5.8	117	5.4	108
10.9 (4.0-12.8)	122	11.1	102	222	12.8	117	11.9	109
20.0 (11.4-32.4)	119	20.1	100	219	23.3	116	21.7	109
49.8 (23.0-60.8)	125	37.9	76	225	41.7 ^b	84 ^b	39.5 ^c	79 ^c
50.0 (19.7-63.0)	121	54.6	109	221	44.9	90	49.8	100
50.0 (11.9-75.9)	123	60.1	120	223	55.8	112	58.0	116

^a All PCB concentrations reported as non-detects by the laboratory.

^b Results excluding the suspect value (results including the suspect value: average = 67.8 ppm and recovery = 136%).

^c Results excluding the suspect value (results including the suspect value: average = 52.8 ppm and Recovery = 106%).

Extract Samples

Percent recovery results for extract samples are summarized in Table 4-6 for the reference laboratory. The average percent recoveries for extract samples ranged from 58 to 121%. In terms of concentration levels, the average recovery at the 10- $\mu\text{g}/\text{mL}$ level (for both solvents) was 104%, compared with 64% at the 100- $\mu\text{g}/\text{mL}$ level. The reference laboratory classified all 16 samples spiked at 10 $\mu\text{g}/\text{mL}$ as Aroclor 1016; however, these samples were actually spiked with Aroclor 1242. Despite this misclassification, the results did not appear to be biased. In contrast, the samples spiked at 100 $\mu\text{g}/\text{mL}$ were correctly classified as Aroclor 1254 but were all biased low. Although these results suggested that Aroclor classification had little effect on the quantification of the extract samples, there was an obvious, consistent error introduced into the analysis of the 100- $\mu\text{g}/\text{mL}$ samples to cause the low bias. For the entire extract data set, the overall percent recovery was 84%.

Table 4-6. Accuracy of the reference laboratory for extract samples

Spike concentration ($\mu\text{g}/\text{mL}$)	Outdoor site			Chamber site			Combined sites	
	Sample ID	Avg conc ($\mu\text{g}/\text{mL}$)	Recovery (%)	Sample ID	Avg conc ($\mu\text{g}/\text{mL}$)	Recovery (%)	Avg conc ($\mu\text{g}/\text{mL}$)	Recovery (%)
0 ^a	129	0	n/a	229	0	n/a	0	n/a
0 ^a	132	0	n/a	232	0	n/a		
10	127	10.9	109	227	9.6	96	10.4	104
10	130	12.1	121	230	8.9	89		
100	128	67.4	67	228	65.2	65	63.5	64
100	131	63.8	64	231	57.7	58		

^a All PCB concentrations reported as non-detects by the laboratory.

Representativeness

Representativeness expresses the degree to which sample data accurately and precisely represent the capability of the method. Representativeness of the method was assessed based on the data generated for clean-QC samples (i.e., method blanks and laboratory control samples) and PE samples. Based on the data assessment (discussed in detail in various parts of this section), it was determined that the representativeness of the reference laboratory data was acceptable. In addition, acceptable performance on laboratory audits substantiated that the data set was representative of the capabilities of the method. In all cases, the performance of the reference laboratory met all requirements for both audits and QC analyses.

Completeness

Completeness is defined as the percentage of measurements that are judged to be usable (i.e., the result was not rejected). Usable results were obtained for 248 of the 256 samples submitted for analysis by the reference laboratory. Eight results (for sample IDs 110 and 112) were deemed incomplete and therefore not valid because the measurements were not quantitative. To calculate completeness, the total number of complete results were divided by the total number of samples submitted for analysis, and then multiplied by 100 to express as a percentage. The completeness of the reference laboratory was 97%, where a completeness of 95% or better is typically considered acceptable.

Comparability

Comparability refers to the confidence with which one data set can be compared with another. The demonstration study was designed to have a one-to-one, sample-by-sample comparison of the PCB results obtained by the reference laboratory and the PCB results obtained by the technology being evaluated. Based on thorough examination of the data and acceptable results on the PE samples, it was concluded that the reference laboratory's SOPs for extraction and analysis, and the data generated using these procedures, were of acceptable quality for comparison with the field technology results. Additional information on comparability was available because the experimental design incorporated randomized analysis of blind, replicate samples. Evaluation of the replicate data implicated some of the individual data points as suspect (see Table D-2). The reference laboratory's suspect data were compared with the technology data on a special-case basis, and exceptions were noted.

Summary of Observations

Table 4-7 provides a summary of the performance of the reference laboratory for the analysis of all sample types used in the technology demonstration study. As shown in Table 4-7, the precision of the PE soils was comparable to the environmental soils. A weighted average, based on the number of samples, gave a best-case precision of 21% and a worst-case precision of 28% for all the soil data (PE and environmental). The extract samples had a smaller overall RSD of 14%. Evaluation of overall accuracy was based on samples with certified or known spiked concentrations (i.e., PE and extract samples). The overall

Table 4-7. Summary of the reference laboratory performance

Sample matrix	Sample type	Number of samples	Precision (average % RSD)	Accuracy (average % recovery)
Blank	Soil	8	n/a ^a	All samples were reported as non-detects.
	Extract	16		
Environmental soil with interferences	Sample ID 110	4	n/a ^a	All samples were reported as non-detects.
	Sample ID 112	4		
Soil Best case (excluding suspect data)	PE	63	18	101
	Environmental			
	< 125 ppm	107	23	n/a ^b
	> 125 ppm	17	19	n/a ^b
	overall	187	21	101
Soil Worst case (including suspect data)	PE	64	21	105
	Environmental			
	< 125 ppm	108	26	n/a ^b
	> 125 ppm	20	56	n/a ^b
	overall	192	28	105
Extract	10 ppm	16	19	104
	100 ppm	16	8	64
	overall	32	14	84

^a Because the results were reported as non-detects, precision assessment is not applicable.

^b Accuracy assessment calculated for samples of known concentration only.

accuracy, based on percent recovery, for the PE samples was 105% for the worst case (which included the suspect value) and 101% for the best case (which excluded the suspect value). These results indicated that the reference laboratory measured values were unbiased estimates of the certified PE concentrations (for samples that contained ≤ 50 ppm of PCBs). Accuracy for the extract samples at 10 ppm was also unbiased, with an average percent recovery of 104%. However, the accuracy for the extract samples at 100 ppm was biased low, with an average recovery of 64%. Overall, the average percent recovery for all extract samples was 84%. The reference laboratory correctly reported all blank samples as non-detects but had difficulty with two soil sample IDs (110 and 112) that contained chemical interferences. In general, the reference laboratory's completeness would be reduced, at the expense of an improvement in precision and accuracy, if the suspect measurements were excluded from the data analysis. Based on this analysis, it was concluded that the reference laboratory results were acceptable for comparison with the developer's technology.

Section 5

Technology Performance and Evaluation

Objective and Approach

The purpose of this section is to present the evaluation of the data generated by the EnviroGard PCB test kit. The technology's precision and accuracy performance are presented for the data generated in the demonstration study. In addition, an evaluation of comparability, through a one-to-one comparison with the reference laboratory data, is presented. An evaluation of other aspects of the technology (such as cost, sample throughput, hazardous waste generation, and logistical operation) is also presented in this section.

Interval Reporting

The EnviroGard results were reported as concentration ranges that were designated as intervals incorporating parentheses/bracket notation. The parentheses indicated that the end-points of the concentration range were excluded, while brackets indicated that the end-points were included. As shown in Table 5-1, the interval (5, 10] indicates that the PCB concentration range is greater than 5 and less than or equal to 10.

As discussed briefly in Section 2 of this report, this technology cannot distinguish between different Aroclors. The test kit has been calibrated to respond in a one-to-one ratio with Aroclor 1248. If site history or information indicates that a different Aroclor (e.g., 1260, 1242, or 1254) is present in the samples, a conversion can be applied (see Table 2-3 for more information). The Aroclor-specific reporting intervals for the EnviroGard results are listed in Table 5-1. For the purposes of the demonstration, SDI was provided information about the type of Aroclor present in the samples. Dilution of samples during analysis to optimize method performance altered some of the standard intervals shown in Table 5-1 for select samples.

Data Assessment

Objective

The purpose of the data assessment section is to present the evaluation of the performance of SDI's EnviroGard PCB test kit through a statistical analysis of the data. PARCC parameters were used to evaluate the test kit's ability to measure PCBs in PE soil, environmental soil, and extract samples. The developer analyzed splits of replicate samples that were also analyzed by the reference laboratory (72 PE soil samples, 136 environmental soil samples, and 24 extract samples). See Section 4 for a more detailed analysis of the reference laboratory's results. Replicate samples were analyzed by the developer at two different sites (under outdoor conditions and inside an environmentally controlled chamber) to evaluate the effect of environmental conditions on the test kit's performance; see Section 3 for further details on the different sites. Evaluation of the measurements made at each site indicated that there were no significant differences between the two data sets. Because environmental conditions did not appear to affect the results significantly, data from both sites were also combined for each parameter (precision and

Table 5-1. EnviroGard PCB test kit reporting intervals

Default mode		Conversion to specific Aroclor			
Interval	Aroclor 1248 concentration range	Interval	Aroclor 1242/1260 concentration range	Interval	Aroclor 1254 concentration range
[0, 1)	$0 \leq \text{PCB ppm} < 1$	[0, 2)	$0 \leq \text{PCB ppm} < 2$	[0, 1.1)	$0 \leq \text{PCB ppm} < 1.1$
[1, 5]	$1 \leq \text{PCB ppm} \leq 5$	[2, 10]	$2 \leq \text{PCB ppm} \leq 10$	[1.1, 5.5]	$1.1 \leq \text{PCB ppm} \leq 5.5$
(5, 10]	$5 < \text{PCB ppm} \leq 10$	(10, 20]	$10 < \text{PCB ppm} \leq 20$	(5.5, 11.1]	$5.5 < \text{PCB ppm} \leq 11.1$
(10, 50]	$10 < \text{PCB ppm} \leq 50$	(20, 100]	$20 < \text{PCB ppm} \leq 100$	(11.1, 55]	$11.1 < \text{PCB ppm} \leq 55$
(50, ∞)	$\text{PCB ppm} > 50$	(100, ∞)	$\text{PCB ppm} > 100$	(55, ∞)	$\text{PCB ppm} > 55$

accuracy) to determine the test kit's overall performance. All statistical tests were performed at the 5% significance level. Appendix D contains the raw data that were used to assess the performance of the EnviroGard test kit.

Precision

Precision is the reproducibility of measurements under a given set of conditions. The frequency with which the same interval was reported within a set of replicates was used to quantify precision. Examples of how the precision was classified are presented in Table 5-2. Reporting a higher number of replicates in the same interval for a given replicate set indicates higher precision. In other words, reporting all four replicate results as the same interval indicates the highest possible precision.

Performance Evaluation Samples

Table 5-3 summarizes the precision information for the EnviroGard test kit's analysis of the PE samples. The EnviroGard test kit reported all four replicates as the same interval (i.e., high precision) for three PE sample sets under each set of environmental conditions (i.e., outdoor and chamber conditions). Operating under the outdoor conditions, seven of eight replicate sets were classified as having medium to high precision. None of the replicate sets was reported with the lowest precision (i.e., none). Under the chamber conditions, medium to high precision was achieved for five of eight replicate sets, with the remaining three replicate sets classified as having low precision. A more detailed analysis of the data showed that the replicates having medium to low precision classifications were never more than one interval away from the most frequently reported interval.

Table 5-2. Classification of precision results

If the replicate results are...	...and the number reported in identical intervals are...	...then the precision classification is...
[0, 1), [1, 5], (5, 10], (10, 50]	0	none
[0, 1), [1, 5], [1, 5], (5, 10]	2	low
[0, 1), [1, 5], [1, 5], [1, 5]	3	medium
[1, 5], [1, 5], [1, 5], [1, 5]	4	high

Table 5-3. Precision of the EnviroGard PCB test kit for PE soil samples

Certified PE Conc. (ppm)	Outdoor site					Chamber site				
	Sample ID	precision				Sample ID	precision			
		none	-----		high		none	-----		high
		Number of replicates reported in identical intervals					Number of replicates reported in identical intervals			
0 ^a	2	3	4	0 ^a	2	3	4			
0	126 ^b				x	226 ^b				x
2.0	118			x		218			x	
2.0	124				x	224				x
5.0	120		x			220		x		
10.9	122				x	222		x		
20.0	119			x		219				x
49.8	125			x		225			x	
50.0	121				x	221				x
50.0	123			x		223		x		
# in each precision classification		0	1	4	3		0	3	2	3

^a Indicates that all four replicates were reported as different intervals.

^b Blank data were not included in the determination of the overall precision.

Environmental Soil Samples

The EnviroGard results for the replicate environmental soil sample measurements are presented in Table 5-4. Under the outdoor conditions, 6 of 17 replicate sets achieved the highest precision classification (i.e., the same interval was reported for all 4 replicates). Under the chamber conditions, 10 of 17 sample sets were classified as high precision. Of the sample sets for which precision was classified as medium to low, only Sample ID 115 had one replicate result that differed by more than one interval range.

Because the majority of measurements fell below 125 ppm, precision was also assessed by partitioning the results into two ranges: low (reference laboratory values < 125 ppm) and high concentrations (reference laboratory values > 125 ppm). See Section 4 for the delineation of which Sample IDs were in the low and high categories. For the low concentrations, 40% of the sample sets were reported with all four replicates in the same interval (i.e., highest possible precision). For the high concentration category, 100% of the sample sets (five of five) were reported with the highest possible precision.

Table 5-4. Precision of the EnviroGard PCB test kit for environmental soil samples

Outdoor site					Chamber site				
Sample ID	precision				Sample ID	precision			
	none	-----		high		none	-----		high
	Number of replicates reported in identical intervals					Number of replicates reported in identical intervals			
	0 ^a	2	3	4		0 ^a	2	3	4
101			x		206				x
102				x	207				x
103			x		208		x		
104			x		209			x	
105				x	210				x
106				x	211				x
107		x			212			x	
108		x			213				x
109		x			214			x	
110		x			215			x	
111			x		216				x
112				x	217				x
113^b				x	201			x	
114		x			202			x	
115			x		203				x
116			x		204				x
117				x	205				x
# in each precision classification	0	5	6	6		0	1	6	10

^a Indicates that all four replicates were reported as different intervals.

^b Bold sample IDs were matching Paducah sample pairs (i.e., 113/201, 114/202, 115/203, 116/204, 117/205).

The Paducah soils (indicated by bold Sample IDs in Table 5-4) were analyzed at both sites to provide an assessment of EnviroGard’s performance under different environmental conditions. For these samples, the data generated under both environmental conditions were also combined to provide an overall assessment of precision. Sample IDs 113 and 201, 114 and 202, 115 and 203, 116 and 204, and 117 and 205 represented replicate Paducah soil sample sets, where the 100 series were samples analyzed under the outdoor conditions, and the 200 series were samples analyzed inside the chamber. Additional statistical analysis was used to compare the effect of the two environmental conditions on the measurements. Results from this analysis showed that there were no significant differences in the data generated at each site. This indicated that these different environmental conditions did not impact the performance of the test kit.

Extract Samples

The EnviroGard results for the replicate extract measurements are presented in Table 5-5. All three sample sets analyzed under the outdoor conditions were reported with the highest possible precision (i.e., all four replicates within the same interval). One sample set (the blank) analyzed under the chamber conditions achieved the highest precision, and the remaining two sample sets were reported with medium precision (i.e., three replicates were reported in the same interval).

Precision Summary

A summary of the EnviroGard test kit's overall precision is presented by sample type (PE, environmental soil, and extract samples) in Table 5-6. For PE and environmental soil samples, 38% and 47% of the samples, respectively, achieved the highest possible precision (i.e., all four sample replicates were reported as the same interval). For the extract samples, 50% of the samples achieved the highest precision.

Table 5-5. Precision of the EnviroGard PCB test kit for extract samples

Outdoor site					Chamber site				
Sample ID	precision				Sample ID	precision			
	none	-----		high		none	-----		high
	Number of replicates reported in identical intervals					Number of replicates reported in identical intervals			
	0 ^a	2	3	4		0 ^a	2	3	4
130 ^b				x	230 ^b				x
131				x	231			x	
132				x	232			x	
# in each precision classification	0	0	0	3		0	0	2	1

^a Indicates that all four replicates were reported as different intervals.

^b Blank data were not included in the determination of the overall precision

Table 5-6. Overall precision of the EnviroGard PCB test kit for all sample types

Environmental site	Percentage of samples classified in each precision category											
	PE samples				Environmental soil samples				Extract samples			
	none	low	med	high	none	low	med	high	none	low	med	high
Outdoor site	0	13	50	38	0	29	35	35	0	0	0	100
Chamber site	0	38	25	38	0	6	35	59	0	0	100	0
Combined sites	0	25	38	38	0	18	35	47	0	0	50	50

Accuracy

Accuracy represents the closeness of the EnviroGard test kit's measured PCB concentrations to the certified values. Because the EnviroGard test kit produced interval results, accuracy was evaluated in terms of the percentage of samples that agreed with, were above (i.e., biased high), and were below (i.e., biased low) the certified value.

Performance Evaluation Soil Samples

Table 5-7 contains a comparison between the EnviroGard interval result and the corresponding certified PE value. The interval(s) listed under a particular column indicates how many of the four replicates were reported as that interval. For example, for Sample ID 120, two replicates were reported as (11.1, 55], and two were reported as (5.5, 11.1]. For Sample ID 119, three are reported as (10, 50], and one is reported as (50, ∞). Note that performance acceptance ranges for the PE results, which are the guidelines established by the provider of the PE materials to gauge acceptable analytical results, are also presented in Table 5-7 for information. These ranges were not used to evaluate the EnviroGard results because the acceptance ranges overlap several EnviroGard reporting intervals.

Table 5-7. EnviroGard test kit accuracy data for PE soil samples

Certified conc. (ppm) (acceptance range, ppm)	Outdoor site					Chamber site				
	Sample ID	# of replicates reported at each interval				Sample ID	# of replicates reported at each interval			
		1	2	3	4		1	2	3	4
0 (n/a)	126				[0, 1)	226				[0, 1)
2.0 (0.7-2.2)	118	(5, 10]		[1, 5]		218	(5, 10]		[1, 5]	
2.0 (0.9-2.5)	124				[2, 10]	224				[2, 10]
5.0 (2.1-6.2)	120		(5.5, 11.1] (11.1, 55]			220	[1.1, 5.5] (5, 10] (5.5, 11.1] (11.1, 55]			
10.9 (4.0-12.8)	122				(20, 100]	222	[2, 10] (20, 100]	(10, 20]		
20.0 (11.4-32.4)	119	(50, ∞)		(10, 50]		219				(10, 50]
49.8 (23.0-60.8)	125	(20, 100]		(55, ∞)		225	(20, 100] (100, ∞)	(55, ∞)		
50.0 (19.7-63.0)	121				(55, ∞)	221				(55, ∞)
50.0 (11.9-75.9)	123	(20, 100]		(100, ∞)		223		(20, 100] (100, ∞)		

The data in Table 5-7 were used to derive the accuracy results that are presented in Table 5-8. Accuracy was based on a comparison of the certified PE value with the interval reported by the EnviroGard test kit. If the interval encompassed the certified PE value, the EnviroGard result “agreed” with the certified value. If the EnviroGard result was above the certified value, the result was classified as “biased high.” If the EnviroGard result was below the certified value, the result was classified as “biased low.” For example, for Sample ID 222, the certified value was 10.9 ppm (for Aroclor 1260). The comparison would be classified as “agreed” for the EnviroGard interval result (10, 20], as “biased high” for the interval result (20, 100], or as “biased low” for the interval result [2, 10]. Separate comparisons were made for the two environmental conditions to determine if ambient temperature and humidity had an effect on the technology performance. Statistical analysis showed that there was no significant difference between the results obtained by the test kit under the two different environmental conditions evaluated in this demonstration. Therefore, all PE sample results were combined to determine the overall percentage of agreement between EnviroGard results and the certified PE value. The overall percentage of agreement was 51%. A single EnviroGard sample result was biased low. A total of 47% of the test kit’s results were biased high, and in approximately 80% of these samples, the difference between the certified PE value and the lower end of the reporting interval was less than 10 ppm. This indicated that many of the results that were biased high were reported in the next highest reporting interval. For example, for PE samples where the certified PE value was 50 ppm, the EnviroGard test kit would generally report the sample concentration as (55, ∞). The large number of sample results that were biased high relative to the certified PE value relates to the conservatism that the manufacturer intentionally incorporates into the EnviroGard’s calculation of reported results.

Table 5-8. Evaluation of agreement between EnviroGard’s PE sample results and the certified PE values as a measure of accuracy

Environmental site	Relative to certified values for performance evaluation samples			Number of samples
	Biased low	Agree	Biased high	
Outdoor site	0%	44%	56%	36
Chamber site	3%	58%	39%	36
Combined sites	1%	51%	47%	72

Extract Samples

Table 5-9 contains a comparison between the EnviroGard interval result and the corresponding spike concentration for the extract samples. The EnviroGard test kit’s percentage of agreement with the spike concentration of the extract samples is summarized in Table 5-10. Statistical analysis showed that environmental conditions had no significant effect upon the performance of the EnviroGard test kit.

Therefore, the data sets generated under the outdoor and chamber conditions were combined. Overall, 14 of 24 extract samples (58%) agreed with the spike concentration. Approximately 38% were biased high, where most of the high bias was associated with the 10 ppm extract samples, all of which were reported

Table 5-9. Accuracy of the EnviroGard test kit for extract samples

Spike conc. ($\mu\text{g/mL}$)	Outdoor site					Chamber site				
	Sample ID	# of replicates reported at each interval				Sample ID	# of replicates reported at each interval			
		1	2	3	4		1	2	3	4
0	132				[0, 1)	232	[1, 5]		[0, 1)	
10	130				(20, 100]	230				(20, 100]
100	131				(55, ∞)	231	(11.1, 55]		(55, ∞)	

Table 5-10. Evaluation of agreement between EnviroGard’s extract results and the spike concentration as a measure of accuracy

Environmental site	Relative to spike concentration for extract samples			Number of samples
	Biased low	Agree	Biased high	
Outdoor site	0%	67%	33%	12
Chamber site	8%	50%	42%	12
Combined sites	4%	58%	38%	24

as (20, 100]. Under the outdoor conditions, no measurements were biased low; under the chamber conditions, one measurement was biased low relative to the spike concentration.

False Positive/False Negative Results

A false positive (fp) result [10] is one in which the technology detects PCBs in the sample when there actually are none. A false negative (fn) result [10] is one in which the technology indicates that there are no PCBs present in the sample, when there actually are. Both fp and fn results are influenced by the method detection limit of the technology. All of the eight blank soil samples were reported as the lowest reporting interval, which included zero, so the fp result was 0%. Of the 192 non-blank soil samples analyzed, EnviroGard reported seven in the lowest reporting interval (e.g., 0 to 2 ppm). All of the corresponding reference laboratory results fell into EnviroGard’s reporting interval (e.g., 1.5 ppm). Therefore, the fn result for the soil samples was 0%. For the eight extract samples, the EnviroGard test kit reported one blank as [1, 5]. Therefore, the fp result was 13%. All other extract samples were reported as non-blanks, so the fn result was 0%.

Representativeness

Representativeness expresses the degree to which the sample data accurately and precisely represent the capability of the technology. The performance data were accepted as being representative of the technology because the EnviroGard test kit was capable of analyzing diverse sample types (PE samples, simulated wipe extract samples, and actual field environmental samples) under multiple environmental conditions. When this technology is used, quality control samples should be analyzed to assess the performance of the EnviroGard PCB test kit under the testing conditions.

Completeness

Completeness is defined as the percentage of measurements that are judged to be useable (i.e., the result was not rejected). Valid results were obtained by the technology for all 232 samples. Therefore, completeness was 100%.

Comparability

Comparability refers to the confidence with which one data set can be compared to another. A one-to-one sample comparison of the EnviroGard results and the reference laboratory results was performed for all soil samples. Accuracy was evaluated in terms of the percentage of samples that agreed with, were above (i.e., biased high), and were below (i.e., biased low) the certified value. For comparability, the EnviroGard results were compared with the results generated by the reference laboratory, including both environmental soils and PE samples. Sample IDs 110 and 112 were excluded because the reference laboratory did not generate quantitative results for these samples. The results are summarized in Table 5-11. The percentage of EnviroGard results that agreed with the reference laboratory results was 53%. Approximately 45% were biased high, and approximately 2% were biased low relative to the results reported by the reference laboratory.

Table 5-11. Evaluation of agreement between EnviroGard’s soil results and the reference laboratory’s results as a measure of comparability

Environmental site	Relative to reference laboratory results for soil samples			Number of samples
	Biased low	Agree	Biased high	
Outdoor site	2%	52%	46%	96
Chamber site	2%	54%	44%	104
Combined sites	2%	53%	45%	200

For the extract samples, the comparison of the EnviroGard test kit’s result with the reference laboratory result was similar to the comparison with the spike concentrations (shown in Table 5-10). There was 63% agreement between the laboratory method and the field technology, and for the rest of the samples (37%), the EnviroGard results were biased high

The soil data not included in previous comparability evaluations (because the replicate data for the reference laboratory were considered suspect) are shown in Table 5-12. Refer to Section 4, in particular Table 4-1, for more information on the reference laboratory’s suspect measurements. The reference laboratory’s suspect data were compared with the EnviroGard’s matching results. For Sample IDs 110 and 112, the reference laboratory obtained qualitative results only. The EnviroGard test kit also had some difficulty with Sample ID 110, producing results in two different intervals, in contrast to Sample ID 112, where all four replicates were reported as the same interval. For the other five suspect values for the reference laboratory data, the EnviroGard test kit generated results that agreed with the replicate means of the reference laboratory. These comparisons demonstrated that the EnviroGard test kit did not have difficulty with most of the samples that were troublesome for the reference laboratory.

Table 5-12. Comparison of the EnviroGard results with the reference laboratory’s suspect measurements

Sample ID	Reference laboratory		EnviroGard	
	Suspect measurement (ppm)	Replicate mean ^a (ppm)	Suspect-matching result (ppm)	Number of replicates reported as this interval
110	≤RDL ^b	≤RDL ^b	(11.1, 55] (55, ∞)	2 2
112	≤RDL ^b	≤RDL ^b	(55, ∞)	4
106	649.6	281.0	(55, ∞)	4
205	3,305.0	493.0	(50, ∞)	4
216	151.6	55.1	(55, ∞)	4
217	1,913.3	659.8	(55, ∞)	4
225	146.0	41.7	(20, 100]	1

^a Mean result excluding the suspect measurement.

^b Measurement reported qualitatively as less than or equal to the reporting detection limit for all replicates.

Summary of PARCC Parameters

Table 5-13 summarizes the EnviroGard test kit’s performance for precision, accuracy, and comparability. Precision was assessed by the percentage of replicate samples where the highest precision was achieved (i.e., all four replicates were reported as the same interval), which was 38% for the PE samples, 47% for the environmental soils, and 50% for the extract samples. The EnviroGard test kit’s agreement and disagreement with certified values were based on the certified PE values (i.e., accuracy) and the reference laboratory results (i.e., comparability). Overall, the EnviroGard test kit’s performance was similar for all samples, because the percentages of agreement and disagreement were not significantly different for each sample type. The percentage in agreement ranged from 48 to 58, the percentage biased high was 38 to 48, and the percentage biased low was 0 to 4.

Table 5-13. EnviroGard PCB test kit performance for precision, accuracy, and comparability

Sample type	Precision ^a	Accuracy ^b			Comparability ^c		
	% high precision	% biased low	% agreed	% biased high	% biased low	% agreed	% biased high
PE	38	1	51	47	4	48	48
Environmental Soil	47	n/a ^b	n/a	n/a	1	56	44
Extract	50	4	58	38	0	63	38

^a Percentage of sample sets that achieved highest precision (i.e., all four replicates were reported as the same interval).

^b EnviroGard result versus certified value; accuracy cannot be assessed for environmental soils.

^c EnviroGard result versus reference laboratory result.

Regulatory Decision-Making Applicability

One objective of this demonstration was to assess the technology's ability to perform at regulatory decision-making levels for PCBs, specifically 50 ppm for soils and 100 µg/100cm² for surface wipes. To assess this, the EnviroGard test kit's performance for PE and environmental soil samples ranging in concentration from 40 to 60 ppm (as determined by the paired reference laboratory analyses) can be used. For this concentration range, the test kit's results agreed with the reference laboratory's results 39% of the time. Results were biased high 59% of the time, and 2% of the results were biased low. No fns were observed for this concentration range. As discussed in the "Accuracy" section, the absence of fns and the presence of a large percentage of results that were biased high were most likely a result of the degree of conservatism that the manufacturer has incorporated into the calculation of results.

Assuming a 10-mL extract volume, extract samples (at 10 and 100 µg/mL) represented surface wipe sample concentrations of 100 µg/100cm² and 1000 µg/100cm². For simulated wipe extract samples, the percentage of the EnviroGard's measurements that agreed with the reference laboratory results was 63%. Approximately 38% of the results were biased high and none were biased low. No fn results were observed for the extract samples.

Additional Performance Factors

Sample Throughput

Sample throughput is representative of the estimated amount of time required to extract the PCBs, to perform appropriate reactions, and to analyze the sample. Operating under the outdoor conditions, the SDI team's sample throughput rate was 18 samples/hour. Working in the chamber, the rate was lower, around 9 to 10 samples/hour. This increased sample throughput under the outdoor conditions may be attributed to the analysis order: because SDI analyzed samples under the chamber conditions first, they may have gained valuable experience that was applied during the analysis of the outdoor samples. Alternatively, SDI may have had more difficulty with the sample matrices that were analyzed only under the outdoor conditions.

Cost Assessment

The purpose of this economic analysis is to provide an estimation of the range of costs for an analysis of PCB-contaminated soil samples using the EnviroGard PCB test kit and a conventional analytical reference laboratory method. The analysis was based on the results and experience gained from this demonstration, costs provided by SDI, and representative costs provided by the reference analytical laboratories that offered to analyze these samples. To account for the variability in cost data and assumptions, the economic analysis was presented as a list of cost elements and a range of costs for sample analysis by the EnviroGard test kit and by the reference laboratory.

Several factors affected the cost of analysis. Where possible, these factors were addressed so that decision-makers can independently complete a site-specific economic analysis to suit their needs. The following categories are considered in the estimate:

- sample shipment costs,
- labor costs,

- equipment costs,
- waste disposal costs.

Each of these cost factors is defined and discussed and serves as the basis for the estimated cost ranges presented in Table 5-14. This analysis assumed that the individuals performing the analyses were fully trained to operate the technology. SDI recommends that new users attend a training session that SDI offers on the use of the EnviroGard test kit. Costs for sample acquisition and pre-analytical sample preparation, which are tasks common to both methods, were not included here.

Table 5-14. Estimated analytical costs for PCB soil samples

EnviroGard PCB test kit Strategic Diagnostics Inc.		EPA SW-846 Method 8080/8081/8082 Reference laboratory	
Sample throughput rate: 18 samples/hour (outdoors) 9–10 samples/hour (chamber)		Typical turn-around time: 14–30 days	
Cost category	Cost (\$)	Cost category	Cost (\$)
Sample shipment	0	Sample shipment	
		Labor	100–200
		Overnight shipping charges	50–150
Labor		Labor	
Mobilization/demobilization	250–400	Mobilization/demobilization	included ^a
Travel	15–1,000 per analyst	Travel	included
Per diem	0–150 per day per analyst	Per diem	included
Rate	30–75 per hour per analyst	Rate	44–239 per sample
Equipment		Equipment	
Mobilization/demobilization	0–150	Mobilization/demobilization	included
Kit rental fee	450 per week	Rental/purchase of system	included
Kit purchase price	2,495	Reagents/supplies	included
Training	< 935		
Reagents/supplies	26 per sample		
Waste disposal	75–1,060	Waste disposal	included

^a“Included” indicates that the cost is included in the labor rate.

EnviroGard Costs

Because the samples were analyzed on site, no sample shipment charges were associated with the cost of operating the EnviroGard test kit. Labor costs included mobilization/demobilization, travel, per diem, and on-site labor.

- Labor mobilization/demobilization: This cost element included the time for one person to prepare for and travel to each site. The estimate ranged from 5 to 8 h, at a rate of \$50 per hour.

-
- **Travel:** This element was the cost for the analyst(s) to travel to the site. If the analyst is located near the site, the cost of commuting to the site (estimated to be 50 miles at \$0.30 per mile) would be minimal (\$15). The estimated cost of an analyst traveling to the site for this demonstration (\$1000) included the cost of airline travel and rental car fees.
 - **Per diem:** This cost element included food, lodging, and incidental expenses and was estimated ranging from zero (for a local site) to \$150 per day per analyst.
 - **Rate:** The cost of the on-site labor was estimated at a rate of \$30 to \$75 per hour, depending on the required expertise level of the analyst. This cost element included the labor involved with the entire analytical process comprising sample preparation, sample management, analysis, and reporting.

Equipment costs included mobilization/demobilization, rental fees or purchase of equipment, and the reagents and other consumable supplies necessary to complete the analysis.

- **Equipment mobilization/demobilization:** This included the cost of shipping the equipment to the test site. If the site is local, the cost would be zero. For this demonstration, the cost of shipping equipment and supplies was estimated at \$150.
- **Rental/purchase:** The fee to rent the EnviroGard test kit at the time of the demonstration was \$450 per week. At the time of the demonstration, the cost of purchasing the equipment was \$2495. The purchase price included a photometer and accessories.
- **Reagents/supplies:** These items are consumable and are purchased on a per-sample basis. At the time of the demonstration, the cost of the reagents and supplies needed to prepare and analyze PCB soil samples using the EnviroGard was \$26 per sample. This cost included the sample preparation supplies, assay supplies, and consumable reagents.

Waste disposal costs are estimated based on the 1997 regulations for disposal of PCB-contaminated waste. The EnviroGard test kit generated approximately 20 lb of vials containing soils and liquid solvents (classified as solid PCB waste suitable for disposal by incineration) and approximately 20 lb of other solid PCB waste (i.e., used and unused soil, gloves, paper towels, ampules, etc.). The cost of disposing of PCB solid waste by incineration at a commercial facility was estimated at \$1.50 per pound. The cost for solid PCB waste disposal at ETPP was estimated at \$18 per pound. The test kit also generated approximately 19 lb of liquid waste. The cost for liquid PCB waste disposal at a commercial facility was estimated at \$0.25 per pound, while the cost at ETPP was estimated at \$11 per pound.

Reference Laboratory Costs

Sample shipment costs to the reference laboratory included the overnight shipping charges, as well as labor charges associated with the various organizations involved in the shipping process.

- **Labor:** This cost element included all of the tasks associated with the shipment of the samples to the reference laboratory. Tasks included packing the shipping coolers, completing the chain-of-custody documentation, and completing the shipping forms. Because the samples contained PCBs, the coolers

were inspected by qualified personnel to ensure compliance with the U.S. Department of Transportation's shipping regulations for PCBs. The estimate to complete this task ranged from 2 to 4 h at \$50 per hour.

- **Overnight shipping:** The overnight express shipping service cost was estimated to be \$50 for one 50-lb cooler of samples.

The labor bids from commercial analytical reference laboratories that offered to perform the PCB analysis for this demonstration ranged from \$44 to \$239 per sample. The bid was dependent on many factors, including the perceived difficulty of the sample matrix, the current workload of the laboratory, and the competitiveness of the market. In this case, the wide range of bids may also be related to the cost of PCB waste disposal in a particular laboratory's state. LAS Laboratories was awarded the contract to complete the analysis as the lowest qualified bidder (\$44 per sample). This rate was a fully-loaded analytical cost, including equipment, labor, waste disposal, and report preparation.

Cost Assessment Summary

An overall cost estimate for the EnviroGard test kit versus the reference laboratory was not made because of the extent of variation in the different cost factors, as outlined in Table 5-14. The overall costs for the application of each technology will be based on the number of samples requiring analysis, the sample type, and the site location and characteristics. Decision-making factors, such as turn-around time for results, must also be weighed against the cost estimate to determine the value of the field technology versus the reference laboratory.

General Observations

The following are general observations regarding the field operation and performance of the EnviroGard test kit:

- The system was light, easily transportable, and rugged. It took about an hour for the SDI team to prepare to analyze samples on the first day of testing. While working at the outdoor site, the SDI team completely disassembled their work station, bringing everything inside, at the close of each day. It took the SDI team less than an hour each morning to prepare for sample analyses.
- Three operators were used for the demonstration because of the number of samples and working conditions, but the technology can be operated by a single person. With three SDI technologies (D TECH, EnviroGard, and RaPID Assay) being demonstrated, SDI elected to work as a team to complete the analyses for each technology (as opposed to three SDI people working with three different technologies).
- Operators generally require 2 to 4 h of training. They should have a basic knowledge of field analytical techniques.
- The SDI team calibrated the photometer often, analyzing 1, 5, 10, and 50 ppm standards and a negative control with every batch of 12 samples. This was done to account for changing environmental conditions (i.e., temperature and humidity).

-
- Data processing and interpretation was minimal. The results were quantified relative to the four calibration standards and reported in terms of intervals using the sample information provided (Aroclor type and ratio).
 - New start and stop solutions were used with every 12 samples. All reagents were allowed to come to room temperature before use. Although it is recommended that all of the reagents in the test kit be stored under refrigerated conditions, the SDI team noted that the reagents can be stored at ambient conditions for several hours.
 - The measurement system (photometer) was battery-operated.
 - The EnviroGard test kit generated approximately 20 lb of vials containing soils and liquid solvents (classified as solid PCB waste suitable for disposal by incineration) and approximately 20 lb of other solid PCB waste (i.e., used and unused soil, gloves, paper towels, ampules, etc.). The test kit also generated approximately 19 lb of liquid waste (aqueous with trace methanol).

Performance Summary

A summary of the performance characteristics of SDI's EnviroGard PCB test kit, presented previously in this chapter, is shown in Table 5-15. The performance of the EnviroGard test kit was characterized as biased, because nearly half (49%) of the EnviroGard results disagreed with the certified PE values, and imprecise, because over half (62%) of the PE replicate results were not reported as the same interval. The test kit had no fp or fn results for the soil samples. For extract samples, the test kit had one fp result and no fn results. It should also be noted that there was an increased likelihood that results would be biased high as a result of the conservatism that the manufacturer has incorporated into the calculation of results.

Table 5-15. Performance summary for the EnviroGard PCB test kit

Feature/parameter	Performance summary								
Blank results	Soils: Correctly reported all 8 samples as [0,1) ppm Extracts: 7 samples reported correctly as [0,1) ppm; 1 sample reported as [1, 5]								
Precision	Percentage of combined sample sets where all four replicates were reported as the same interval PE soils: 38% Environmental soils: 47% Extracts: 50%								
Accuracy	<table border="0" style="width: 100%;"> <tr> <td style="width: 50%;">PE soils</td> <td style="width: 50%;">Extracts</td> </tr> <tr> <td>agreed = 51%</td> <td>agreed = 58%</td> </tr> <tr> <td>biased high = 47%</td> <td>biased high = 38%</td> </tr> <tr> <td>biased low = 1%</td> <td>biased low = 1%</td> </tr> </table>	PE soils	Extracts	agreed = 51%	agreed = 58%	biased high = 47%	biased high = 38%	biased low = 1%	biased low = 1%
PE soils	Extracts								
agreed = 51%	agreed = 58%								
biased high = 47%	biased high = 38%								
biased low = 1%	biased low = 1%								
False positive results	Blank soils: 0% (0 of 2 samples) Blank extracts: 13% (1 of 8 samples)								
False negative results	PE and environmental soils: 0% (0 of 192 samples) Spiked extracts: 0% (0 of 16 samples)								
Comparison with reference laboratory results	<table border="0" style="width: 100%;"> <tr> <td style="width: 50%;">PE and environmental soils</td> <td style="width: 50%;">Extracts</td> </tr> <tr> <td>agreed = 53%</td> <td>agreed = 63%</td> </tr> <tr> <td>biased high = 45%</td> <td>biased high = 38%</td> </tr> <tr> <td>biased low = 2%</td> <td>biased low = 0%</td> </tr> </table>	PE and environmental soils	Extracts	agreed = 53%	agreed = 63%	biased high = 45%	biased high = 38%	biased low = 2%	biased low = 0%
PE and environmental soils	Extracts								
agreed = 53%	agreed = 63%								
biased high = 45%	biased high = 38%								
biased low = 2%	biased low = 0%								
Regulatory decision-making applicability	<table border="0" style="width: 100%;"> <tr> <td style="width: 50%;">PE and environmental soils (40 to 60 ppm)</td> <td style="width: 50%;">Extracts (100 µg/100cm², 1000 µg/100cm²)</td> </tr> <tr> <td>agreed = 39%</td> <td>agreed = 63%</td> </tr> <tr> <td>biased high = 59%</td> <td>biased high = 38%</td> </tr> <tr> <td>biased low = 2%</td> <td>biased low = 0%</td> </tr> </table>	PE and environmental soils (40 to 60 ppm)	Extracts (100 µg/100cm², 1000 µg/100cm²)	agreed = 39%	agreed = 63%	biased high = 59%	biased high = 38%	biased low = 2%	biased low = 0%
PE and environmental soils (40 to 60 ppm)	Extracts (100 µg/100cm², 1000 µg/100cm²)								
agreed = 39%	agreed = 63%								
biased high = 59%	biased high = 38%								
biased low = 2%	biased low = 0%								
Sample throughput	9–10 samples/hour (chamber) 18 samples/hour (outdoors)								
Power requirements	Battery-operated photometer								
Operator requirements	Basic knowledge of chemical techniques; 2–4 hours technology-specific training								
Cost	Incremental: \$26 per sample Instrumental: \$2,495 (purchase); \$450 (weekly rental)								
Hazardous waste generation	Approximately 20 lb of solid/liquid (classified as solid PCB waste suitable for disposal by incineration) Approximately 20 lb of solid (used gloves, pipettes, paper towels, etc.) Approximately 19 lb of liquid waste (aqueous with trace methanol)								

Section 6

Technology Update and Representative Applications

Objective

The purpose of this section is to allow the developer to provide information regarding new developments with its technology since the demonstration activities. In addition, the developer has provided a list of representative applications in which its technology has been or is currently being used.

Technology Update

Reconfiguration of Soil Extraction (Sample Preparation) Products

SDI is in the process of commercializing a common extraction kit for three of four remediation immunoassay test kit product lines. The affected product lines include the EnviroGard, EnSys (not demonstrated here), and RaPID Assay test kit systems. The new “Universal Extraction Kit” will be used with assay kits of these three product lines, with extraction solvents or dilution reagents specifically formulated to match individual kits available as kit component options where required. The new test kit configuration will provide increased user convenience and simplify the product specification and ordering process without affecting test kit analytical performance. Commercialization of the new Universal Extraction Kit was initiated in April 1998. The new kits are not for use with the D TECH product line, which will continue to use the existing SDI Soil Extraction Pac products.

Instrument Consolidation

Associated with the incorporation of several independently developed product lines into SDI’s product offerings, some consolidation of equipment and instrumentation is anticipated in the near future. This will consist primarily of reducing the number of pipette types and photometers used to perform the assays. While pipette types and procedures for pipetting reagents and reading and interpreting assay results may change slightly, no effect on assay performance will result.

Representative Applications

In a 1997 report entitled *Field Analytical and Site Characterization Technologies: Summary of Applications* [11], the use of SDI immunoassay kits is documented at more than 30 remediation sites under state or federal oversight. Contact information is provided for many of the immunoassay kit users at these sites. The summary report can be obtained from the National Center for Environmental Publications and Information (NCEPI). Hard copies of the report can be ordered, free of charge, by telephone, (513) 891-6561; by fax (513) 891-6685; or through the NCEPI homepage on the Internet at <http://www.epa.gov/ncepihom/>. The summary report is available for viewing or downloading as a .pdf file from the CLU-IN Internet website. The Internet address is <http://clu-in.com/pubichar.htm>.

Data Quality Objective Example

This application of SDI's EnviroGard PCB immunoassay kit is based on data quality objective (DQO) methods for project planning advocated by the American Society for Testing and Materials [12, 13] and EPA [14]. ORNL derived a DQO example from the performance results in Section 5. This example, which is presented in Appendix E, illustrates the use of the performance data for SDI's EnviroGard from the ETV demonstration in the DQO process to select the number of samples to characterize the decision rule's fp and fn error rates.

Section 7

References

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- [11] U.S. Environmental Protection Agency. *Field Analytical and Site Characterization Technologies: Summary of Applications*, EPA-542-R-97-011, Office of Solid Waste and Emergency Response, Washington, D.C., November 1997

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- [12] American Society for Testing and Materials (ASTM). *Standard Practice for Generation of Environmental Data Related to Waste Management Activities Quality Assurance and Quality Control Planning and Implementation*, D5283-92, 1997.
- [13] American Society for Testing and Materials (ASTM), *Standard Practice for Generation of Environmental Data Related to Waste Management Activities Development of Data Quality Objectives*, D5792-95, 1997.
- [14] U.S. Environmental Protection Agency. *Guidance for Data Quality Assessment*, EPA QA/G-9; EPA/600/R-96/084, EPA, Washington, D.C., July 1996.

Appendix A
Description of Environmental Soil Samples



Table A-1. Summary of soil sample descriptions

Location	Request for Disposal (RFD) #	Drum #	Description
Oak Ridge	40022	02	Soil from spill cleanup at the Y-12 Plant in Oak Ridge, Tennessee. This soil is PCB-contaminated soil excavated in 1992.
Oak Ridge	40267	01 02 03 04	Soil from the Elza Gate area, a DOE Formerly Utilized Sites Remedial Action Program site in Oak Ridge, Tennessee. This soil is PCB-contaminated soil that was excavated in 1992.
Oak Ridge	24375	01 02 03	Catch-basin sediment from the K-711 area (old Powerhouse Area) at the DOE East Tennessee Technology Park (formerly known as Oak Ridge Gaseous Diffusion Plant) in Oak Ridge, Tennessee. This soil is PCB-contaminated storm drain sediment that was excavated in 1991.
Oak Ridge	43275	01 02	Soil from the K-25 Building area at the DOE East Tennessee Technology Park (formerly known as Oak Ridge Gaseous Diffusion Plant) in Oak Ridge, Tennessee. This soil is PCB-contaminated soil that was excavated in 1993.
Oak Ridge	134555	03	Soil from the K-707 area at the DOE East Tennessee Technology Park (formerly known as Oak Ridge Gaseous Diffusion Plant) in Oak Ridge, Tennessee. This soil is PCB-contaminated soil from a dike spillage that was excavated in 1995.
Paducah	97002	01 02 03 04	Soil from the DOE Paducah Gaseous Diffusion Plant in Kentucky. This soil is PCB-contaminated soil from a spill cleanup at the C-746-R (Organic Waste Storage Area) that was excavated in 1989.
Portsmouth	7515	858 1069 1096 1898 2143 2528 3281 538 940 4096	Soil from the DOE Portsmouth Gaseous Diffusion Plant in Ohio. This soil is PCB-contaminated soil from a probable PCB oil spill into the East Drainage Ditch that was excavated in 1986.
Tennessee Reference Soil	n/a	n/a	Captina silt loam from Roane County, Tennessee; used as a blank in this study (i.e., not contaminated with PCBs)



Appendix B
Characterization of Environmental Soil Samples

Table B-1. Summary of environmental soil characterization

Location	Sample ID	RFD drum # ^a	Composition			Total organic carbon (mg/kg)	pH
			% gravel	% sand	% silt + clay		
Oak Ridge	101	40022-02	0	91.8	8.2	5384	7.12
	102	40267-03	0.5	99.3	0.2	13170	7.30
	103	40267-01	0.2	96.7	3.1	13503	7.21
	104	40267-04	0.6	98.2	1.2	15723	7.07
	105	40267-01S ^b	0.5	94.8	4.7	14533	7.28
	106	24375-03	0.5	87.8	11.7	19643	7.36
	107	24375-01	2.5	92.5	5.0	1196	7.26
	108	40267-02	0.4	94.2	5.4	9007	7.30
	109	24375-02	0.3	93.1	6.6	1116	7.48
	110	43275-01	0	89.2	10.8	14250	7.57
	111	134555-03S ^b	0.5	88.1	11.4	10422	7.41
	112	43275-02	0.1	91.4	8.5	38907	7.66
	126, 226	non-PCB soil	0	85.6	14.4	9249	7.33
	Paducah	113, 201	97002-04	0	92.4	7.6	1296
114, 202		97002-01	0.2	87.6	12.2	6097	7.64
115, 203		97002-03	0.1	83.6	16.3	3649	7.59
116, 204		97002-02	0.4	93.7	5.8	4075	7.43
117, 205		97002-02S ^b	0.4	93.7	5.8	4075	7.43
Portsmouth	206	7515-4096	0	87.1	12.9	3465	7.72
	207	7515-1898	0.2	78.0	21.8	3721	7.66
	208	7515-1096	0.4	74.4	25.2	3856	7.77
	209	7515-2143	0	74.3	25.7	10687	7.71
	210	7515-0940	0.3	73.0	26.7	7345	7.78
	216	7515-0538	0.5	73.3	26.3	1328	7.78
	211	7515-0538S ^b					
	217	7515-0538S ^b					
	212	7515-2528	0.5	70.4	29.1	5231	7.92
	213	7515-3281	0.5	72.6	26.8	5862	7.67
	214	7515-0858	0	65.8	34.2	6776	7.85
	215	7515-1069	1.3	75.0	23.7	4875	7.56

^a Request for disposal drum number (see Table A-1).

^b "S" indicates that the environmental soil was spiked with additional PCBs.

Appendix C
Temperature and Relative Humidity Conditions



Table C-1. Average temperature and relative humidity conditions during testing periods

Date	Outdoor site		Chamber site	
	Average temperature (°F)	Average relative humidity (%)	Average temperature (°F)	Average relative humidity (%)
7/22/97	85	62	70 ^a	38 ^a
7/23/97	85	70	60 ^a	58 ^a
7/24/97	85	67	58	66
7/25/97	80	70	56	54
7/26/97	85	55	57	51
7/27/97	80	75	55	49
7/28/97	79	88	57	52
7/29/97	<i>b</i>	<i>b</i>	55	50

^a The chamber was not operating properly on this day. See discussion in Section 3.

^b No developers were working outdoors on this day.

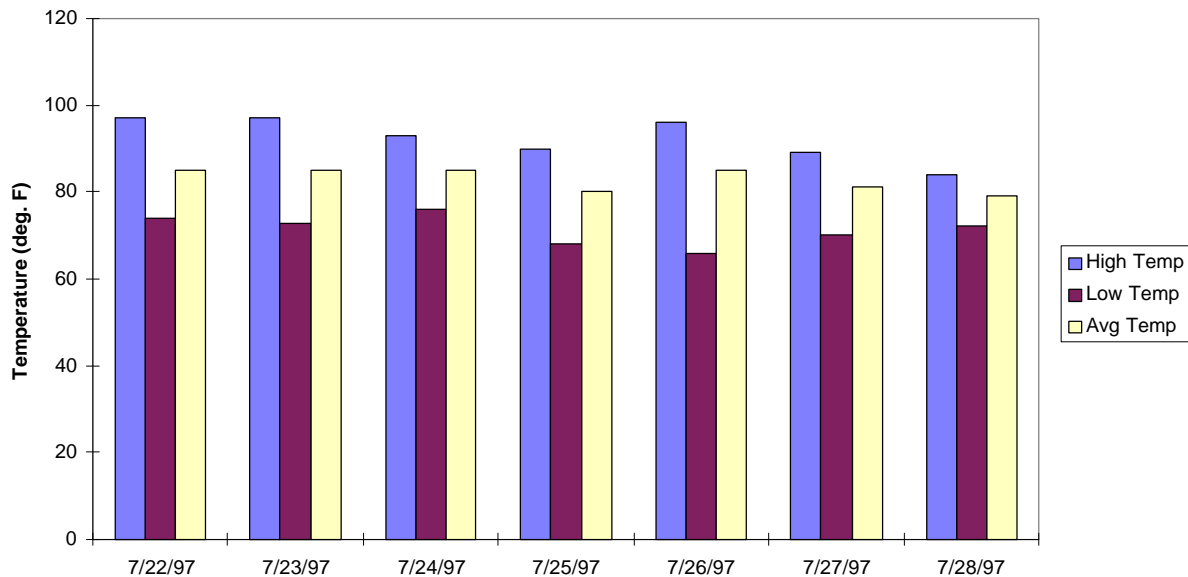


Figure C-1. Summary of temperature conditions for outdoor site.

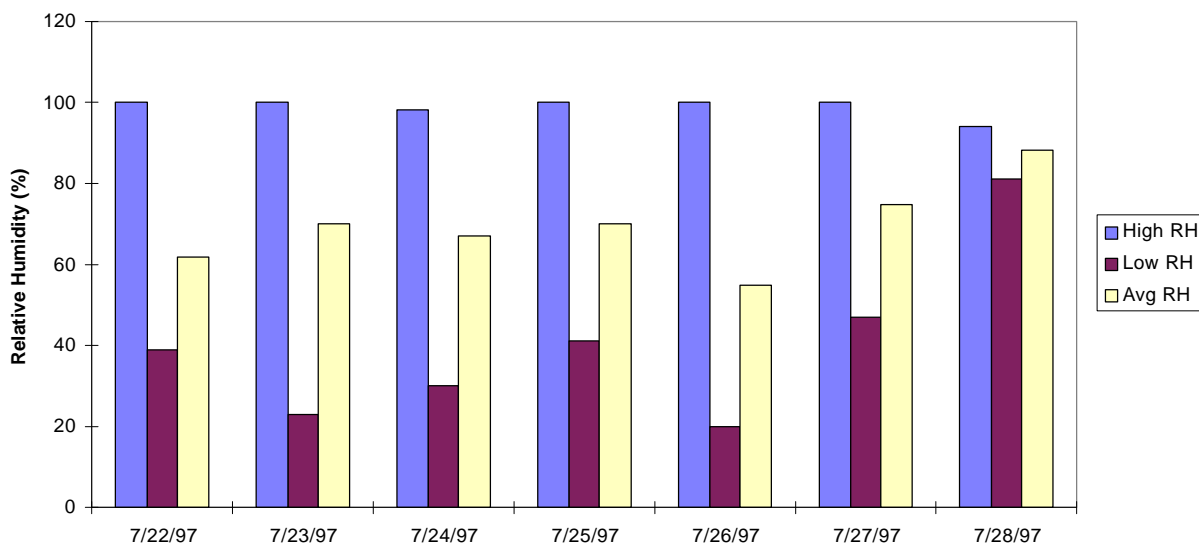


Figure C-2. Summary of relative humidity conditions for the outdoor site.

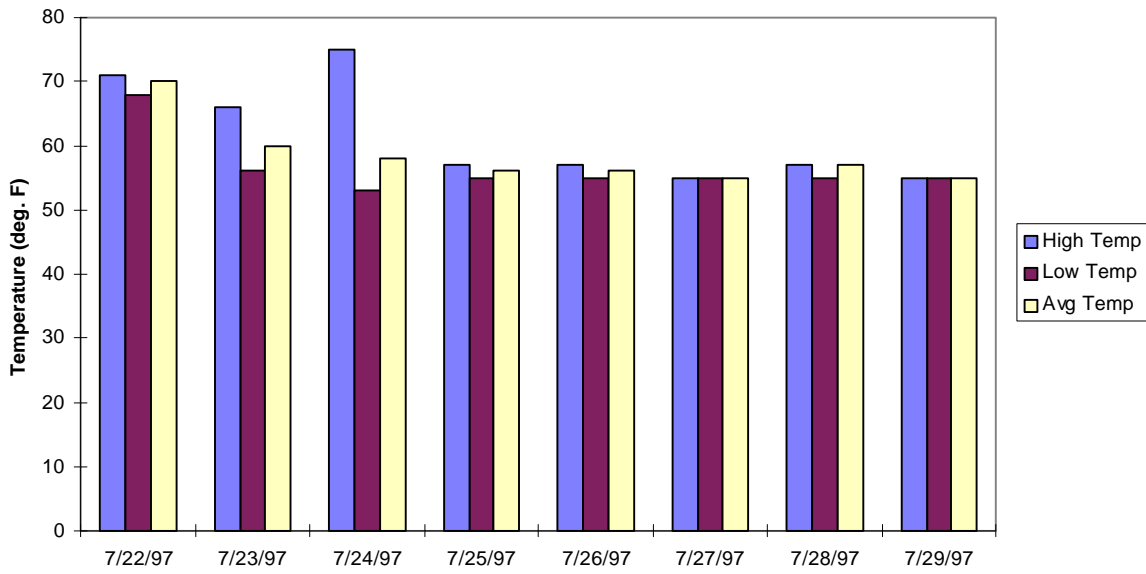


Figure C-3. Summary of temperature conditions for chamber site.

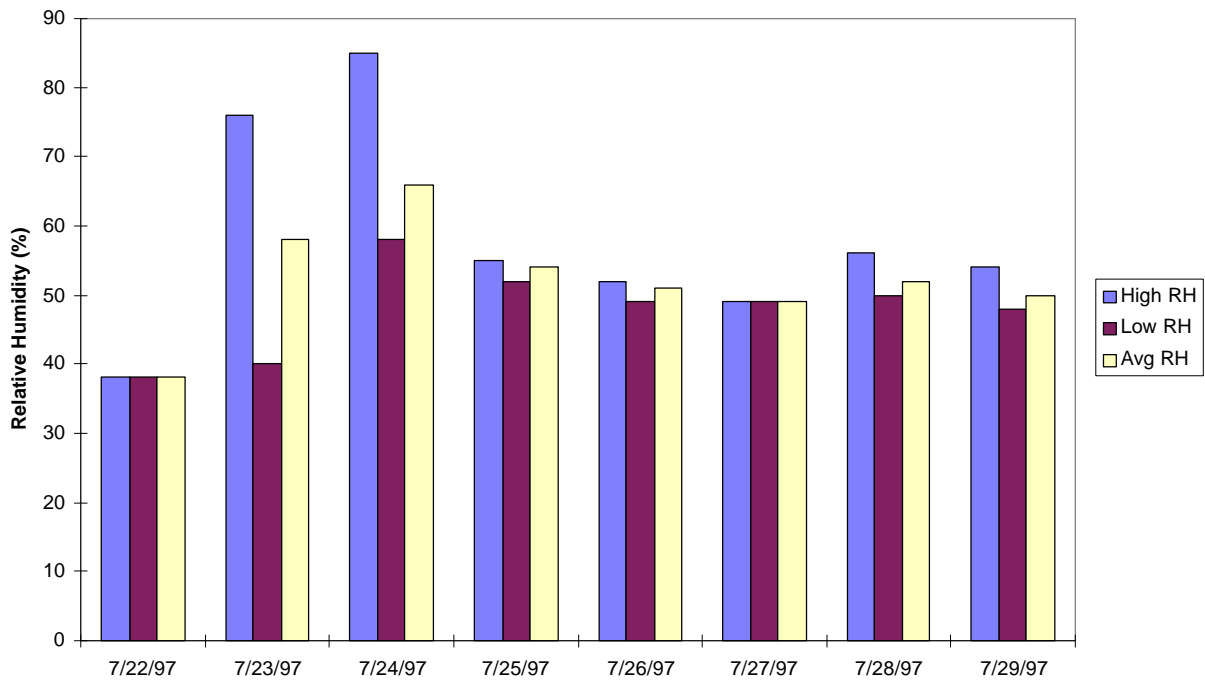


Figure C-4. Summary of relative humidity conditions for chamber site.

Appendix D
EnviroGard PCB Test Kit
PCB Technology Demonstration Sample Data

Legend for Appendix D Tables

Table Heading	Definition
Obs	Observation
Sample ID	Sample identification 101 to 126 = outdoor site soil samples 127 to 130 = outdoor site extract samples 201 to 226 = chamber site soil samples 227 to 230 = chamber site extract samples
Rep	Replicate of sample ID (1 through 4)
EnviroGard Result	EnviroGard's measured PCB concentration (ppm)
Ref lab result	LAS reference laboratory measured PCB concentration (ppm) Values with "≤" are samples that the reference laboratory reported as "≤ reporting detection limit"
Reference Aroclor	Aroclor(s) identified by the reference laboratory
Type	Sample = environmental soil 1242, 1248, 1254, 1260 = Aroclor in PE samples Blank = non-PCB-contaminated sample
Order	Order of sample analysis by SDI (started with 2001–2116, then 1001–1116)

Table D-1. EnviroGard PCB technology demonstration soil sample data

Obs	Sample ID	Rep	EnviroGard Result (ppm)	Ref Lab Result (ppm)	Reference Aroclor	Type	Order
1	101	1	[0,4]	0.6	1254	Sample	1007
2	101	2	[0,4]	0.4	1254	Sample	1009
3	101	3	[0,4]	0.5	1254	Sample	1019
4	101	4	[0,2)	0.5	1254	Sample	1073
5	102	1	[1.1,5.5]	2.2	1254	Sample	1047
6	102	2	[1.1,5.5]	2.1	1254	Sample	1083
7	102	3	[1.1,5.5]	1.7	1260	Sample	1024
8	102	4	[1.1,5.5]	2.5	1260	Sample	1032
9	103	1	(5.5,11.1]	3.0	1254	Sample	1014
10	103	2	(5.5,11.1]	2.4	1254	Sample	1077
11	103	3	[1.1,5.5]	2.0	1260	Sample	1057
12	103	4	(5.5,11.1]	1.6	1260	Sample	1099
13	104	1	(11.1,55]	6.8	1260	Sample	1015
14	104	2	(11.1,55]	6.0	1254	Sample	1045
15	104	3	(11.1,55]	14.8	1254	Sample	1071
16	104	4	(5.5,11.1]	9.9	1254	Sample	1059
17	105	1	(20,100]	49.7	1260	Sample	1055
18	105	2	(20,100]	84.1	1260	Sample	1064
19	105	3	(20,100]	50.6	1260	Sample	1051
20	105	4	(20,100]	53.2	1260	Sample	1063
21	106	1	(55,∞)	269.6	1254	Sample	1094
22	106	2	(55,∞)	255.9	1254	Sample	1049
23	106	3	(55,∞)	317.6	1254	Sample	1087
24	106	4	(55,∞)	649.6	1254	Sample	1036
25	107	1	[1.1,5.5]	1.0	1254	Sample	1025
26	107	2	(5.5,11.1]	1.6	1254	Sample	1101
27	107	3	(5.5,11.1]	1.2	1254	Sample	1102
28	107	4	[1.1,5.5]	1.2	1254	Sample	1044
29	108	1	[1.1,5.5]	1.7	1254	Sample	1040
30	108	2	(5.5,11.1]	2.0	1254	Sample	1097
31	108	3	[1.1,5.5]	1.7	1254	Sample	1089
32	108	4	(5.5,11.1]	1.9	1254	Sample	1021
33	109	1	(5.5,11.1]	1.5	1254	Sample	1017
34	109	2	[1.1,5.5]	2.1	1254	Sample	1037
35	109	3	[1.1,5.5]	1.8	1254	Sample	1016
36	109	4	(5.5,11.1]	2.4	1254	Sample	1002
37	110	1	(55,∞)	≤490.0	Non-Detect	Sample	1028
38	110	2	(55,∞)	≤99.0	Non-Detect	Sample	1034
39	110	3	(11.1,55]	≤66.0	Non-Detect	Sample	1050
40	110	4	(11.1,55]	≤98.0	Non-Detect	Sample	1062

Obs	Sample ID	Rep	EnviroGard Result (ppm)	Ref Lab Result (ppm)	Reference Aroclor	Type	Order
41	111	1	(11.1,55]	44.5	1254	Sample	1069
42	111	2	(11.1,55]	36.0	1254	Sample	1033
43	111	3	(11.1,55]	39.3	1254	Sample	1067
44	111	4	(55,∞)	35.1	1254	Sample	1079
45	112	1	(55,∞)	≤66.0	Non-Detect	Sample	1088
46	112	2	(55,∞)	≤200.0	Non-Detect	Sample	1100
47	112	3	(55,∞)	≤130.0	Non-Detect	Sample	1060
48	112	4	(55,∞)	≤200.0	Non-Detect	Sample	1056
49	113	1	[1.1,5.5]	0.7	1260	Sample	1003
50	113	2	[1.1,5.5]	1.1	1260	Sample	1092
51	113	3	[1.1,5.5]	0.6	1260	Sample	1001
52	113	4	[1.1,5.5]	1.9	1248/1260	Sample	1072
53	114	1	[2,10]	1.1	1260	Sample	1022
54	114	2	[2,10]	1.2	1260	Sample	1048
55	114	3	[0,2)	1.3	1260	Sample	1011
56	114	4	[0,2)	1.7	1260	Sample	1053
57	115	1	(20,100]	14.9	1248	Sample	1054
58	115	2	(20,100]	12.4	1016	Sample	1068
59	115	3	(20,100]	15.0	1248	Sample	1081
60	115	4	[2,10]	16.9	1248	Sample	1093
61	116	1	(20,100]	41.4	1248	Sample	1058
62	116	2	(100,∞)	41.2	1016	Sample	1070
63	116	3	(100,∞)	48.5	1248	Sample	1018
64	116	4	(100,∞)	34.0	1016	Sample	1096
65	117	1	(50,∞)	431.6	1016	Sample	1043
66	117	2	(50,∞)	406.3	1016	Sample	1008
67	117	3	(50,∞)	304.7	1016	Sample	1012
68	117	4	(50,∞)	392.8	1016	Sample	1066
69	118	1	[1,5]	2.1	1248	1248	1039
70	118	2	[1,5]	1.9	1016	1248	1052
71	118	3	(5,10]	0.7	1248	1248	1103
72	118	4	[1,5]	1.6	1248	1248	1004
73	119	1	(10,50]	21.2	1016	1248	1082
74	119	2	(10,50]	17.2	1248	1248	1086
75	119	3	(10,50]	17.4	1248	1248	1035
76	119	4	(50,∞)	24.4	1248	1248	1095
77	120	1	(11.1,55]	4.5	1254	1254	1029
78	120	2	(5.5,11.1]	4.0	1254	1254	1030
79	120	3	(5.5,11.1]	6.3	1254	1254	1076
80	120	4	(11.1,55]	5.0	1254	1254	1042
81	121	1	(55,∞)	58.7	1254	1254	1075
82	121	2	(55,∞)	55.7	1254	1254	1013
83	121	3	(55,∞)	53.2	1254	1254	1031
84	121	4	(55,∞)	50.9	1254	1254	1023
85	122	1	(20,100]	12.2	1260	1260	1041
86	122	2	(20,100]	10.9	1260	1260	1046
87	122	3	(20,100]	11.3	1260	1260	1091
88	122	4	(20,100]	10.0	1260	1260	1038

Obs	Sample ID	Rep	EnviroGard Result (ppm)	Ref Lab Result (ppm)	Reference Aroclor	Type	Order
89	123	1	(100,∞)	59.2	1260	1260	1080
90	123	2	(20,100]	56.9	1260	1260	1061
91	123	3	(100,∞)	66.8	1260	1260	1098
92	123	4	(100,∞)	57.5	1260	1260	1104
93	124	1	[2,10]	1.8	1254	1254/1260	1074
94	124	2	[2,10]	1.4	1260	1254/1260	1006
95	124	3	[2,10]	1.9	1254	1254/1260	1026
96	124	4	[2,10]	1.8	1254	1254/1260	1005
97	125	1	(55,∞)	32.0	1254	1254/1260	1010
98	125	2	(55,∞)	41.3	1254	1254/1260	1027
99	125	3	(20,100]	46.0	1254	1254/1260	1065
100	125	4	(55,∞)	32.2	1260	1254/1260	1020
101	126	1	[0,1)	≤0.1	Non-Detect	Blank	1084
102	126	2	[0,1)	≤0.1	Non-Detect	Blank	1078
103	126	3	[0,1)	≤0.2	Non-Detect	Blank	1085
104	126	4	[0,1)	≤1.3	Non-Detect	Blank	1090
105	201	1	[1.1,5.5]	1.0	1016/1260	Sample	2085
106	201	2	[1.1,5.5]	1.0	1016/1260	Sample	2074
107	201	3	[1,5]	1.1	1016/1260	Sample	2007
108	201	4	[1.1,5.5]	0.6	1260	Sample	2079
109	202	1	[2,10]	1.4	1260	Sample	2012
110	202	2	[2,10]	1.6	1260	Sample	2083
111	202	3	[2,10]	1.2	1260	Sample	2060
112	202	4	[0,2)	1.5	1260	Sample	2045
113	203	1	(20,100]	14.0	1248	Sample	2021
114	203	2	(20,100]	12.8	1248	Sample	2058
115	203	3	(20,100]	16.2	1248	Sample	2075
116	203	4	(20,100]	12.4	1248	Sample	2062
117	204	1	(100,∞)	43.1	1248	Sample	2048
118	204	2	(100,∞)	45.3	1248	Sample	2009
119	204	3	(100,∞)	41.0	1248	Sample	2059
120	204	4	(100,∞)	47.7	1248	Sample	2006
121	205	1	(50,∞)	3305.0	1016/1260	Sample	2072
122	205	2	(50,∞)	538.7	1016	Sample	2041
123	205	3	(50,∞)	457.0	1016	Sample	2024
124	205	4	(50,∞)	483.3	1016	Sample	2047
125	206	1	[2,10]	2.9	1260	Sample	2097
126	206	2	[2,10]	1.1	1260	Sample	2038
127	206	3	[2,10]	1.1	1016/1260	Sample	2050
128	206	4	[2,10]	2.5	1260	Sample	2067
129	207	1	(20,100]	17.8	1260	Sample	2034
130	207	2	(20,100]	14.3	1260	Sample	2020
131	207	3	(20,100]	21.6	1260	Sample	2063
132	207	4	(20,100]	21.6	1254	Sample	2044
133	208	1	(55,∞)	42.0	1260	Sample	2028
134	208	2	(55,∞)	27.7	1016/1260	Sample	2025
135	208	3	(20,100]	24.0	1254	Sample	2013
136	208	4	(20,100]	28.4	1260	Sample	2019

Obs	Sample ID	Rep	EnviroGard Result (ppm)	Ref Lab Result (ppm)	Reference Aroclor	Type	Order
137	209	1	(55,∞)	32.7	1260	Sample	2040
138	209	2	(55,∞)	79.3	1260	Sample	2089
139	209	3	(20,100]	11.0	1260	Sample	2104
140	209	4	(55,∞)	37.9	1260	Sample	2100
141	210	1	(55,∞)	123.2	1260	Sample	2094
142	210	2	(55,∞)	61.5	1260	Sample	2017
143	210	3	(55,∞)	84.1	1260	Sample	2046
144	210	4	(55,∞)	85.5	1260	Sample	2093
145	211	1	(55,∞)	387.8	1254	Sample	2042
146	211	2	(55,∞)	581.4	1254	Sample	2003
147	211	3	(55,∞)	330.0	1254	Sample	2023
148	211	4	(55,∞)	318.7	1254	Sample	2057
149	212	1	(10,20]	3.8	1260	Sample	2031
150	212	2	[2,10]	3.9	1260	Sample	2099
151	212	3	[2,10]	4.3	1260	Sample	2052
152	212	4	[2,10]	0.8	1260	Sample	2069
153	213	1	(20,100]	6.9	1260	Sample	2010
154	213	2	(20,100]	7.3	1260	Sample	2032
155	213	3	(20,100]	7.8	1260	Sample	2005
156	213	4	(20,100]	10.5	1260	Sample	2084
157	214	1	(20,100]	26.0	1260	Sample	2030
158	214	2	(20,100]	25.6	1260	Sample	2103
159	214	3	(20,100]	29.1	1260	Sample	2070
160	214	4	(55,∞)	20.2	1260	Sample	2033
161	215	1	(20,100]	25.1	1260	Sample	2027
162	215	2	(55,∞)	24.1	1260	Sample	2087
163	215	3	(20,100]	26.2	1260	Sample	2066
164	215	4	(20,100]	31.2	1016/1260	Sample	2077
165	216	1	(55,∞)	151.6	1260	Sample	2092
166	216	2	(55,∞)	47.0	1260	Sample	2043
167	216	3	(55,∞)	54.3	1260	Sample	2101
168	216	4	(55,∞)	64.0	1260	Sample	2055
169	217	1	(55,∞)	886.7	1254	Sample	2001
170	217	2	(55,∞)	549.8	1254	Sample	2064
171	217	3	(55,∞)	542.8	1254	Sample	2014
172	217	4	(55,∞)	1913.3	1016/1260	Sample	2095
173	218	1	[1,5]	2.8	1248	1248	2091
174	218	2	[1,5]	2.4	1248	1248	2026
175	218	3	(5,10]	2.6	1248	1248	2098
176	218	4	[1,5]	2.6	1248	1248	2061
177	219	1	(10,50]	22.4	1248	1248	2073
178	219	2	(10,50]	26.0	1016	1248	2015
179	219	3	(10,50]	29.4	1248	1248	2036
180	219	4	(10,50]	15.2	1248	1248	2082
181	220	1	(5,10]	8.5	1254	1254	2035
182	220	2	[1.1,5.5]	4.9	1254	1254	2039
183	220	3	(5.5,11.1]	4.7	1254	1254	2054
184	220	4	(11.1,55]	5.2	1254	1254	2090

Obs	Sample ID	Rep	EnviroGard Result (ppm)	Ref Lab Result (ppm)	Reference Aroclor	Type	Order
185	221	1	(55,∞)	32.0	1016/1260	1254	2080
186	221	2	(55,∞)	44.1	1016/1260	1254	2018
187	221	3	(55,∞)	43.8	1254	1254	2076
188	221	4	(55,∞)	59.6	1254	1254	2096
189	222	1	[2,10]	13.2	1260	1260	2051
190	222	2	(20,100]	12.4	1260	1260	2088
191	222	3	(10,20]	12.7	1260	1260	2002
192	222	4	(10,20]	12.7	1260	1260	2049
193	223	1	(20,100]	56.6	1260	1260	2071
194	223	2	(100,∞)	50.3	1260	1260	2029
195	223	3	(100,∞)	49.9	1260	1260	2008
196	223	4	(20,100]	66.4	1260	1260	2011
197	224	1	[2,10]	2.2	1254	1254/1260	2056
198	224	2	[2,10]	1.2	1260	1254/1260	2065
199	224	3	[2,10]	1.4	1260	1254/1260	2086
200	224	4	[2,10]	2.1	1254	1254/1260	2078
201	225	1	(55,∞)	56.4	1260	1254/1260	2081
202	225	2	(55,∞)	36.5	1016/1260	1254/1260	2004
203	225	3	(100,∞)	32.1	1260	1254/1260	2068
204	225	4	(20,100]	146.0	1254	1254/1260	2022
205	226	1	[0,1)	≤0.1	Non-Detect	Blank	2037
206	226	2	[0,1)	≤0.8	Non-Detect	Blank	2053
207	226	3	[0,1)	≤0.1	Non-Detect	Blank	2102
208	226	4	[0,1)	≤0.1	Non-Detect	Blank	2016

Table D-2. EnviroGard PCB test kit technology demonstration extract sample data

Obs	Sample ID	Rep	EnviroGard Result (ppm)	Ref Lab Result (ppm)	Reference Aroclor	Type	Spike ^a (ppm)	Order
1	130	1	(20,100]	16.4	1016	1242	10	1113
2	130	2	(20,100]	10.9	1016	1242	10	1110
3	130	3	(20,100]	10.3	1016	1242	10	1109
4	130	4	(20,100]	10.7	1016	1242	10	1114
5	131	1	(55,∞)	67.1	1254	1254	100	1105
6	131	2	(55,∞)	57.1	1254	1254	100	1106
7	131	3	(55,∞)	62.8	1254	1254	100	1112
8	131	4	(55,∞)	68.2	1254	1254	100	1111
9	132	1	[0,1)	≤0.1	Non-Detect	Blank	0	1108
10	132	2	[0,1)	≤0.1	Non-Detect	Blank	0	1107
11	132	3	[0,1)	≤0.1	Non-Detect	Blank	0	1116
12	132	4	[0,1)	≤0.1	Non-Detect	Blank	0	1115
13	230	1	(20,100]	9.8	1016	1242	10	2112
14	230	2	(20,100]	10.4	1016	1242	10	2111
15	230	3	(20,100]	7.6	1016	1242	10	2105
16	230	4	(20,100]	7.9	1016	1242	10	2116
17	231	1	(55,∞)	55.2	1254	1254	100	2109
18	231	2	(11.1,55]	55.0	1254	1254	100	2115
19	231	3	(55,∞)	61.3	1254	1254	100	2108
20	231	4	(55,∞)	59.1	1254	1254	100	2114
21	232	1	[0,1)	≤0.1	Non-Detect	Blank	0	2106
22	232	2	[1,5]	≤0.1	Non-Detect	Blank	0	2113
23	232	3	[0,1)	≤0.1	Non-Detect	Blank	0	2107
24	232	4	[0,1)	≤0.1	Non-Detect	Blank	0	2110

^aNominal spike concentration of the extract sample prepared by ORNL.

Table D-3. Corrected reference laboratory data

Error	Sample ID	Reported Result (ppm)	Corrected Result (ppm)
Transcription	106	≤490	255.9
	130	5.6	10.3
	205	32,000	3,305.0
	207	180	17.8
	210	160	123.2
Calculation	118	3.6	2.1
	119	4.3	17.4
	209	2.3	37.9
	214	43.0	26.0
	219	29.0	22.4
Interpretation	101 ^a	≤0.7	0.5
	101 ^a	≤0.7	0.6
	107	≤1.3	1.2
	109	18.0	1.5
	113 ^b	≤0.9	0.6
	113 ^b	≤1.0	0.7
	119	18.0	21.2
	127	7.2	10.9
	201	≤ 1.0	0.6
	219	21.0	26.0

^a Two of four measurements in Sample ID 101 were corrected.

^b Two of four measurements in Sample ID 113 were corrected.

Appendix E
Data Quality Objective Example

Disclaimer

The following hypothetical example serves to demonstrate how the information provided in this report may be used in the data quality objectives (DQO) process. This example serves to illustrate the application of quantitative DQOs to a decision process, but it cannot attempt to provide a thorough education in this topic. Please refer to other educational or technical resources for further details. Additionally, since the focus of this report is on the analytical technology, this example makes the simplifying assumption that the contents of these drums will be homogeneous. In the real world, however, this assumption is seldom valid, and matrix heterogeneity constitutes a source of considerable uncertainty which must be adequately evaluated if the overall certainty of a site decision is to be quantified.

Background and Problem Statement

An industrial company discovered a land area contaminated with PCBs from an unknown source. The contaminated soil was excavated into waste drums. Preliminary characterization determined that a number of PCB drums had to be incinerated to reduce or eliminate the PCB contamination. The incinerated soil was placed in drums for disposal in a landfill. However, a final check of each drum was required to verify for the regulator that the appropriate level of cleanup had been achieved. The regulator required that no drum have more than 2 ppm of PCB. The company's DQO team was considering the use of the EnviroGard PCB test kit to measure the PCB concentration in each drum. Because the type of Aroclor was unknown, all measurements would be reported as Aroclor 1254. The plan was to randomly select soil samples collected from each drum and test with the kit to determine if the concentration was in one of the three intervals: $[0, 1.1)$, $[1.1, 5.5]$, or $(5.5, \infty)$. Recall that this notation describes the concentration ranges $0 \text{ ppm} \leq \text{PCB} < 1.1 \text{ ppm}$, $1.1 \text{ ppm} \leq \text{PCB} \leq 5.5 \text{ ppm}$, and $\text{PCB} > 5.5 \text{ ppm}$, as used in Section 5. The DQO team decided that a drum would be reprocessed by incineration if any of the EnviroGard results indicated a concentration in the intervals $[1.1, 5.5]$, or $(5.5, \infty)$. In agreement with the regulator, the DQO team determined that a decision rule for disposal would be based on the number of samples with PCB concentrations in the intervals $[1.1, 5.5]$, or $(5.5, \infty)$.

General Decision Rule

If all of the PCB sample results show concentrations in $[0, 1.1)$, then send the soil drum to the landfill.

If any of the PCB sample results are different from $[0, 1.1)$ then reprocess the soil drum by incineration.

DQO Goals

EPA's *Guidance for Data Quality Assessment* [14] states in Sect. 1.2: "The true condition that occurs with the more severe decision error...should be defined as the null hypothesis." The DQO Team decided that the more severe decision error would be for a drum to be erroneously sent to a landfill if the drum's PCB concentration actually exceeded the 2 ppm limit. Therefore, the null hypothesis is constructed to assume that a drum's true PCB concentration exceeds the 2 ppm limit, and as a "hot" drum, it would be sent to the incinerator. Drums would be sent to the landfill only if the null hypothesis is rejected and it is concluded that the "true" average PCB concentration is less than 2 ppm.

With the null hypothesis defined in this way, a false positive (fp) decision is made when it is concluded that a drum contains less than 2 ppm PCBs (i.e., the null hypothesis is rejected), when actually the drum is “hot” (i.e., the null hypothesis is true). The team required that the error rate for sending a “hot” drum to the landfill (i.e., the fp error rate for the decision) could not be more than 5%. Therefore, a sufficient number of samples must be taken from each drum that the fp decision error rate (FP) is 0.05 (or less) if the true drum concentration is 2 ppm. This scenario represents a 5% chance of sending a drum containing 2 ppm or more of PCBs to the landfill. The EnviroGard interval boundary of 1.1 ppm can be used as a conservative estimate of the 2 ppm criterion.

The DQO team did not want to send an excessive number of drums to the incinerator if the average PCB concentration was less than 2 ppm because of the expense. In this situation, a false negative (fn) decision is made when it is concluded that a drum is “hot” (i.e., the null hypothesis is not rejected), when in actuality, the drum contains soil with less than 2 ppm PCBs (i.e., the null hypothesis is actually false). After considering the guidelines presented in Section 1.1 of EPA’s *Guidance for Data Quality Assessment* [14], the DQO team recommended that the fn decision error rate (FN) be 0.10 if the true drum concentration was less than 1.1 ppm. That is, there would be a 10% probability of sending a drum to the incinerator (denoted as Pr[Take Drum to Incinerator]) if the true PCB concentration for a drum was less than 1.1 ppm.

Permissible FP and FN Error Rates and Critical Decision Point

FP: $\text{Pr}[\text{Take Drum to Landfill}] \leq 0.05$ when true PCB concentration ≥ 1.1 ppm

FN: $\text{Pr}[\text{Reprocess Drum in Incinerator}] \leq 0.10$ when true PCB concentration < 1.1 ppm

Use of Technology Performance Information to Implement the Decision Rule

Technology performance information is used to evaluate whether a particular analytical technology can produce data of sufficient quality to support the site decision. Because the DQO team is considering the use of the EnviroGard PCB kit, the performance of this technology [as reported in this Environmental Technology Verification (ETV) report] was used to assess its applicability to this project. The question arises: How many samples are needed from a single drum to permit a statistically valid decision at the specified certainty? Recall that the simplifying assumption was made that the PCB distribution throughout the soil within a single drum is homogeneous and thus matrix heterogeneity will not contribute to overall variability. The only variability, then, to be considered in this example is the variability in performance of the EnviroGard kit’s analytical method, which is determined by precision and accuracy studies.

Determining the Number of Samples

The number of samples needed to satisfy the FP and FN requirements depends on the misclassification error rates of the EnviroGard PCB kit. Two types of misclassifications have to be considered:

1. Underestimating the PCB concentration—classifying a sample concentration in $[0, 1.1)$ when the true PCB concentration is greater or equal to 1.1 ppm

-
2. Overestimating the PCB concentration—classifying a sample concentration in [1.1, 5.5] or (5.5, ∞) when the PCB concentration is less than 1.1 ppm.

The ETV demonstration results on performance evaluation soil samples and on environmental soil samples indicated the error rates for the two types of misclassifications to be

$$P_U = \text{Pr}[\text{Underestimating the PCB concentration}] = 0.017$$

$$P_O = \text{Pr}[\text{Overestimating the PCB concentration}] = 0.400$$

The probability distribution of classifying the number of soil samples in different concentration intervals follows a binomial probability distribution [7]. This probability distribution and the requirements for FP and FN can be used to determine the number of samples to meet the DQO goals. The FP for the decision rule is related to P_U by

$$FP = \text{Pr}[\text{All EnviroGard results} < 1.1 \text{ ppm for } PCB \geq 1.1 \text{ ppm}] = (P_U)^n \quad (\text{E-1})$$

The FP error rate decreases as the sample size increases. The sample size is solved as

$$n = \frac{\text{Log}(FP)}{\text{Log}(P_U)} \quad (\text{E-2})$$

where

- n = number of samples from a drum to be measured
- FP = false positive decision error rate (e.g., FP = 0.05)
- P_U = probability of underestimating the PCB concentration

$$n = \frac{\text{Log}(0.05)}{\text{Log}(0.017)} = \frac{-1.301}{-1.770} = 0.74 \quad 1 \text{ .}$$

The sample size was rounded up to the next integer, an operation that will decrease the FP for the decision rule. The DQO team would have to analyze one sample from each drum to meet the decision rule's fp requirement. The FN for the decision rule is related to P_O by

$$FN = \text{Pr}[\text{Some of EnviroGard results} \geq 1.1 \text{ ppm for } PCB < 1.1 \text{ ppm}] = 1 - (1 - P_O)^n \quad (\text{E-3})$$

The probability of an fn decision (FN = sending a drum for reprocessing) actually increases with increasing sample size because the chance that the kit will overestimate a result increases with continued testing. The sample size required to meet the FN requirement is

$$n = \frac{\text{Log}(1 - FN)}{\text{Log}(1 - P_o)} \quad (\text{E-4})$$

where

- n = number of samples from a drum to be measured
- FN = false negative decision error rate (e.g., FN = 0.10)
- P_o = probability of overestimating a PCB concentration

$$n = \frac{\text{Log}(1 - 0.10)}{\text{Log}(1 - 0.400)} = \frac{-0.046}{-0.222} = 0.21 \quad .$$

The sample size must be rounded up to n = 1 (fractions of a sample analysis are not possible). When n = 1, and the above equation is solved for FN, it is found that the DQO team cannot meet its goal of 10% FN and would have to accept an FN of 40%. This situation occurs because of the 40% overestimation error rate of the kit. If a decision about a drum is based on a single sample, and that sample has a 40% chance of being overestimated, there is consequently a 40% chance that the drum will be unnecessarily sent for reprocessing through the incinerator (which was the definition of FN). Although this amount of conservatism may be desirable in some situations, in others it may not be. The only way to reduce the FN in this kind of scenario is to use an analytical technology with a lower overestimation error rate.

The DQO team in our example decided that the sampling procedure would be to randomly select one soil sample from each drum and test the sample with the EnviroGard PCB kit. The DQO team would send the drum to the landfill if the EnviroGard result was less than 1.1 ppm, and send the drum to be reprocessed by incineration if the EnviroGard result was greater than 1.1 ppm. To meet the FP requirement of 5%, the DQO team would have to accept an FN of 40%.

Decision Rule for 5% FP and 40% FN

If one randomly selected soil sample has a PCB test result reported as the interval [0, 1.1) then send the soil drum to the landfill.

If one randomly selected soil sample has a PCB test result different from [0, 1.1) then reprocess the soil drum by incineration.

Alternative FP Parameter

The following statement describes how changing the FP requirement from 5% to 0.1% would affect the decision rule. Using $FP = 0.001$, the calculated sample sizes would be $n = 1.70$, which is rounded up to 2. The FN would be 64%. The higher FN occurs because each of the two samples has a 40% chance of being overestimated, and if only one is overestimated, the drum is sent for reprocessing. The decision rule for the lower FP requirement would be

Decision Rule for $FP = 0.1\%$ and $FN = 64\%$

If two randomly selected soil samples have their PCB test results reported as $[0, 1.1)$ then send the soil drum to the landfill.

If either of the two randomly selected soil samples has a PCB test result different from $[0, 1.1)$ then reprocess the soil drum by incineration.